

Comparison of techniques to localise U-bearing particles in environmental samples

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26	environmental samples							
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41 Abstract

This publication shows results of a comparison of three techniques for localising radioactive, and U-bearing particles on same samples. Particles are localised by the means of three methods: (1) Fission Tracks (FT), (2) Imaging Plate (IP), and (3) real time autoradiography (BeaQuant[®]). These techniques were applied to various samples, including a sediment sampled in

the vicinity of the Fukushima Dai-Ichi Nuclear Power Plant (FDNPP) and a sample made of pure
U oxide particles. In addition, the efficiency of the combination of two methods (FT and IP) to
localise specifically anthropogenic U-bearing particles was tested.

49 Key words

50 BeaQuant[®], imaging plate, fission tracks, alpha and beta emitters localisation, uranium.

51 **1. Introduction**

Nowadays there is a real need for the detection and characterisation of anthropogenic 52 53 radioactive particles contained in soils/sediments to assess the exposure of human beings to 54 ionizing radiation and its consequences on health and the environment. These particles have been 55 emitted during nuclear accidents such as Chernobyl [1–3], Fukushima Dai-Ichi Nuclear Power 56 Plant (FDNPP) [4–16], Thulé [17, 18], and Palomares [19–21] or have been released from nuclear facilities during nuclear fuel cycle operations [22]. Actinides are particularly interesting because 57 of their long persistence in the environment $(24.11 \times 10^3 \text{ years for } {}^{239}\text{Pu to } 4.47 \times 10^9 \text{ years for }$ 58 ²³⁸U) and can therefore be used as a sediment tracer [23]. For accurate assessment of the 59 environmental impact of these particles and associated risks, it is important to be free from other 60 sources of actinides like natural uranium bearing minerals or global fallout from past nuclear 61 62 weapon tests. Thus when performing traditional bulk analysis on soil or sediment samples, the 63 resulting isotopic and elemental composition is a mixing of all these actinide sources. Only particle 64 analysis that consists in the characterisation at the particle scale enables to focus on the emission source of interest. 65

66 The goal of this work is to investigate nuclear fuel-bearing particles and to study the detection capacities of three techniques: (1) the fission track method, based on the fissile properties 67 of 235 U and 239 Pu, (2) real time α -autoradiography, based on the detection of α -emitting isotopes 68 $(^{234}\text{U}, ^{235}\text{U}, ^{238}\text{U}, ^{239}\text{Pu}, ^{240}\text{Pu})$ and (3) α -autoradiography on imaging plates. Since the 1990's, the 69 70 fission track method has been implemented for the detection of particles containing uranium and 71 plutonium in safeguards samples [24] and is very sensitive to high enriched uranium bearing 72 particles. In the field of medical biology and geosciences, a new real-time digital autoradiography method is already in use, the BeaOuant[®] (Ai4R, Nantes, France). This device, implemented on 73 74 thin slides of rocks [25] and on biological samples, enables a quantitative chemical mapping of 75 alpha emitters. The last method based on the use of imaging plates is commonly used in the 76 literature to detect particles containing β - γ emitters [4–6, 8–16, 26, 27]. These β - γ emitters may be actinide decay products or fission products such as ¹³⁷Cs if nuclear fuel has been irradiated. In 77 78 this paper, we implemented the imaging plate method developed by [27] to detect specifically α -79 emitters.

80 To our knowledge, these three detection-localisation techniques are compared for the first 81 time for two different types of samples: (1) pure U oxide micro-particles (highly enriched U with a 40% atomic abundance of ²³⁵U); and (2) a sediment sample collected in the vicinity of FDNPP 82 83 with significant (> 10,000 kBq·kg⁻¹) β and γ emitter activities that may contain nuclear fuel 84 particles. Besides, a grass sample (IAEA 472 CRM) that contains only β and γ emitters was also 85 prepared as a control sample. We also describe the different limitations of each method. Actually, 86 the presence of naturally occurring nuclides such as uranium (α emitter) and its decay products 87 (mainly β - γ emitters) in soils/sediments may hide the presence of anthropogenic particles specifically in case of volcanic U-rich soils for example. Moreover, in the case of irradiated nuclear 88

particles, radioactive emissions from actinides are mostly negligible compared to the proportion of β - γ emissions from fission products (¹³⁷Cs, etc.). The fission products may generate interferences on α -autoradiography detection. In addition to this comparison exercise, theoretical numbers of FT and α -tracks generated for the U-bearing particles from both samples (pure U and FDNPP U-bearing particles) were calculated assuming given transmission and auto-adsorption coefficients.

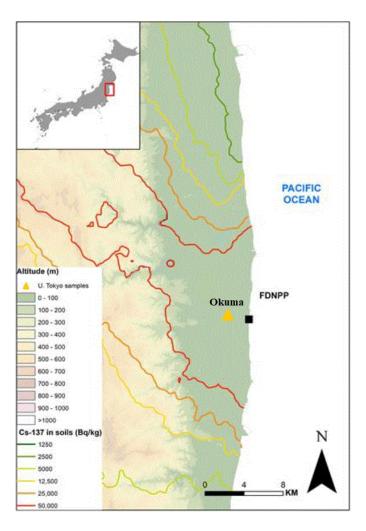
95

2. Material and methods

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2.1. Sample collection and preparation

97 A sediment sample referred to as Okuma was collected at 2.7 km to the West of FDNPP, 98 in Choja-bara, Otto-zawa, Okuma Town, Fukushima Prefecture, Japan (37.42230°N 141.00345°E) (Fig. 1) on March 18, 2014 within the Restricted Zone. The specific activity of ¹³⁷Cs was 13,300 99 $kBq\cdot kg^{-1}$ (± 18 kBq·kg⁻¹) at the time of analysis. About 1 g of the sample was dry-sieved to 63 100 101 µm. The sample fraction was then suspended in a mixture of ethanol and of an organic polymer (collodion) and spread on Lexan[®] disks (Lexan[®], General Electric Plastic, USA). Coating with 102 103 collodion was necessary both to immobilize particles and to prevent contamination of the detection 104 media. Three disks referred to as 20B2, 20B16, and 20B20, were prepared and used for the 105 experiments. A process blank was also prepared with the same method.



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Fig. 1 Background ¹³⁷Cs level map in soils of the Fukushima prefecture after Chartin et al. (2013)
with location of the Fukushima Dai-Ichi Nuclear Power Plant (FDNPP, black square) and the
investigated sediment sample (yellow triangle).

111 Moreover, a subsample of the IAEA-472 certified material was also analysed as a control 112 sample. This material, collected in Polesskoe, Kiev, Ukraine during summer in 1990, is a grass 113 sample, which contains only β and γ emitters like ⁴⁰K and ¹³⁷Cs with a reference weight activity 114 decay-corrected to 2020 of 1 kBq·kg⁻¹ in ⁴⁰K and 8 kBq·kg⁻¹ in ¹³⁷Cs. This standard sample was 115 prepared using the same protocol as for the Okuma sample and dispersed onto two Lexan[®] disks

(1 mg, i.e. $\approx 4.1 \times 10^{-3}$ Bq of ¹³⁷Cs per disk). The analysis of this certified material was used to 116 estimate the contribution of β emissions during BeaQuant[®] acquisitions. 117

118 Finally, samples containing highly enriched uranium micro-particles (HEU MP) with a 119 ²³⁵U atomic abundance of 40 % referred to as 19A32 and 19A33 were also used as a reference for 120 α -emitters because their specific α -activity (α -activity by mass unit) is higher than those of the 121 naturally-occurring (non-purified) U (by a factor of ~6) and of the U from reactor 3 (without Pu) 122 from FDNPP (by a factor of ~ 17). Therefore, we expect to have greater chances to detect such 123 pure HEU particles by α -autoradiography techniques instead of natural mineral particles rich in U 124 for the same U content. These particles, whose equivalent diameters range between 250 and 400 125 nm, were mixed with urban particulate matter (SRM 1648a, National Institute of Standards and 126 Technology, USA) that contains element like aluminium, silicon and lead which are expected to 127 be found in industrial environments. This SRM contains also natural U with an estimated concentration about 5 mg kg^{-1} , which is slightly larger than the estimate of the average 128 concentration of U in soil (~ $3 \text{ mg} \cdot \text{kg}^{-1}$). The estimated mass of natural U was equal to about 16 129 130 ng and the mass of HEU equal to about 32 ng [29]. These particles were trapped in a piece of 131 cotton and were extracted from the cotton tissue in an ultrasonic bath filled with ethanol. After mixing with the organic polymer, particles were deposited on Lexan[®] disks like the Okuma 132 sample. 133

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2.2. Methods for localising radioactive particles implemented in this study

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2.2.1. Imaging Plates (IP) technique

137 The imaging plate (IP) is a passive, two-dimensional radiation detector based on the photo-138 stimulated luminescence (PSL) phenomenon [30, 31]. Due to their high sensitivity over a large

139 range of energy levels, IPs applications are widely used in medical and environmental sciences. 140 There are three types of IPs: (1) TR (Tritium), (2) MS (Multipurpose Standard), and (3) SR (Super 141 Resolution). Differences between these IP types are: (1) the absence of protective layers for the 142 TR type and (2) variable thickness of the sensitive layer between IP types. In all of them, the active layer is made of a photo-stimulating phosphor crystal, typically BaF(Br,I):Eu²⁺ [31]. When 143 144 exposed to an ionizing radiation – energetic charged particles, X-rays, or γ rays – some of the 145 incoming ionizing particles deposit a part of their energy which is stored in the phosphorous layer in the form of FBr⁻ or FI⁻ centres and Eu³⁺ ions [30]. The electrons that are trapped in these sites. 146 can be later released by visible light irradiation and reintegrated into the Eu³⁺ ions followed by the 147 148 emission of 3 eV photons called PhotoStimulated Light (PSL) [30]. The emitted light in the blue 149 part of the visible spectrum can be detected, usually by means of a photomultiplier tube (PMT), 150 and its intensity is proportional to the incident radiation intensity recorded in the IP [33] with a 151 very large dynamic range of intensity, typically spanning five orders of magnitude.

152 To be used as a detector, IP must be calibrated for both the PSL response to radiation (energy and 153 type) and for spatial resolution. The IP calibration is done with known activity sources and 154 different exposure times [34]. In this study, we implemented TR films (IP, TR 12.5 cm \times 25.2 cm, 155 Perkin Elmer) without a protective layer to increase the probability of detecting α emissions at 156 very low levels of activity from actinide-bearing particles. A methodology based on the overlay of 157 several IPs to try to identify different radioisotopes was developed [27]. Indeed, γ rays and 158 electrons can imprint several IPs depending on their energy while α particles can only impact the 159 first (closest to the same sample) IP. Here, we applied this method by overlaying two IPs to 160 discriminate the impact of β and γ versus α -particles. The IP in contact with prepared disks (IP1) 161 allows recording α , β and γ PSL and the second IP (IP2) allows recording of only γ and β PSL.

Therefore, PSL emissions observed only for IP1 and not for IP2, or significantly stronger for IP1
than for IP2, are very likely to indicate that samples contain α-emitters.

164 Two disks of the Fukushima sample (Okuma) and two disks of the HEU MP were analysed 165 with this device for an acquisition time of two weeks in the darkness in lead shielding 166 (Supplementary information A). This acquisition time was set to improve the chance to detect 167 alpha emission without much loss of sensitivity due to fading effect. Autoradiography images were 168 then recorded using the IP reader (Cyclone Plus, Perkin Elmer) with a pixel size of 150 dpi (dot 169 per inch) corresponding to one measurement each 169 µm in the IP screen. The IP reader quantified 170 the radiation intensity as PSL for each pixel (a total of 749×1452 pixels) and stored it in a TIFF 171 image format.

To identify high-radioactivity spots, the autoradiography images of IP1 (**Fig. 4a**) and IP2 (**Fig. 4b**) were processed using the ImageJ software (<u>https://imagej.nih.gov/ij/</u>) in order to report the PSL values of each IP pixel in a table. Raw PSL values were corrected by the software's drift values of the IP reader (Eq. (1)).

176
$$PSL_{Corrected IP1} = PSL^2 \times 2.34 \times 10^{-5} (1)$$

For the IP2 the same calculation takes into account the loss factor of β-γ between the two IP, which determined to be 0.18 from a ¹³⁷Cs source [27] (Eq. (2)).

179
$$PSL_{corrected IP2} = PSL^2 \times 2.34 \times 10^{-5} / 0.18 (2)$$

180 To highlight hot spots corresponding to α emissions only, the PSL-corrected values of the IP2 were 181 subtracted from the PSL-corrected values of the IP1 (Eq. (3)).

$$PSL_{1-2} = PSL_{corrected IP1} - PSL_{corrected IP2} (3)$$

183 Finally, *PSL*₁₋₂ was corrected for the background and the standard deviation (Eq. (4))

184
$$PSL_{1-2} = PSL_{1-2} - mean_{background} - 3 \times RSD_{background}$$
(4)

185 Results of the processed IP PSL were then converted into a new autoradiography image with
186 ImageJ to visualize hot spot locations (Fig. 4c).

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2.2.2. Fission Tracks (FT) technique

An additional Lexan[®] disk, acting as solid state nuclear track detector and referred to as 189 "FT disk", was welded onto each Lexan® disk onto which the particulate sample was deposited 190 191 (referred to as "deposition" disk). The disks were then packed into a capsule and irradiated in the 192 TRIGA nuclear reactor (Jozef Stefan Institute, Ljubljana, Slovenia). A specific irradiation device 193 was used for that purpose in order to get a well-thermalized neutron flux with a high fluence, in the 10^{15} neutron cm⁻² range, to be able to detect particles with low numbers of fissile nuclei [35]. 194 Fissile nuclides (²³⁵U, ²³⁹Pu) underwent a fission reaction during which two fission fragments were 195 emitted. These fragments then impacted the Lexan[®] disks and modified their surface structure [24]. 196 197 After irradiation, the FT disk was chemically etched to make FT visible under an optical 198 microscope. The sea urchin-shaped FT clusters are characteristic of the presence of a fissile 199 nuclide-bearing particle. With appropriate landmarks on both deposition and revelation Lexan[®] 200 disks, it is possible to locate the particle that produced the FT cluster.

FT clusters may come from either (1) a mineral particle of significant size with a sufficiently high content of naturally-occurring U (like zircons, monazite or granites), or (2) a particle coming from the FDNPP which contains human-modified U (i.e. purified and enriched in fissile 235 U isotope) and possibly 239 Pu (produced within the nuclear fuel by neutron capture or initially added to UO₂ to make a MOX fuel). Once the particle localised, its source can be identified by combining the determinations of the elemental composition from SEM-EDS and of the U isotopic composition by secondary ion mass spectrometry.

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2.2.3. BeaQuant[®] technique

The BeaOuant[®] is a digital autoradiography real-time acquisition system based on the use of a 210 211 gaseous detection medium [25, 36]. The micro-pattern gaseous detector (MPGD) incorporates a 212 micromesh parallel ionization multiplier (PIM) [37-39]. In this gas mixture, charged particle 213 emitted by the sample interact with the gas and release their energy by ionization of the gas 214 mixture. Exposed to appropriate electric fields, the electrons created during the interaction are 215 multiplied, drifted and sent to the segment anode. A α -disintegration triggering the electronic acquisition is thus reported on an autoradiography image in real time. BeaQuant[®] has the capability 216 217 to count and map α or β -emitting particles without influence of γ emitters [25, 37] and with a high 218 spatial resolution (maximum resolution of 20 μ m for both α and β from tritium detection [39, 40]. With the MPGD, the BeaQuant's $\$ sensitivity is about 5×10^{-5} cps·cm⁻² for a U concentration of 2 219 220 ppm at secular equilibrium [25], with a detection linear response of 5 orders of magnitude [37]. 221 Moreover, by adjusting the BeaQuant[®] amplification gains, it is possible to discriminate α and β 222 emissions.

However, samples analysed with BeaQuant[®] need to be perfectly flat: any surface 223 224 roughness can induce artefacts corresponding to artificial hot spots on the autoradiography image. The sample preparation carried out in this study is therefore perfectly appropriate for BeaQuant[®] 225 226 analyses. Disks were placed during four days within a GS-1010 sample holder (10 cm \times 10 cm), in front of the gas detector and analysed with the measurement parameters configured for ²³⁸U in 227 228 order to detect α emitters (referred to as method 1) specifically. As Fukushima samples are highly contaminated with 137 Cs (>10 kBq·kg⁻¹), a second acquisition was realized with a configuration 229 more restrictive on U energy in order to avoid ¹³⁷Cs detection (referred to as method 2). The 230

acquisition time was set to ten days because the sensitivity was reduced with these settings. The
IAEA-472 standard sample was also analysed by method 2 in order to check the decrease in
background due to the energy limitation. For method 2, the pixel size chosen for the analysis
reached up to 50 µm and increased to 400 µm after post-processing in order to integrate a stronger
signal per pixel.

After acquisition with both methods, post-processing was conducted to remove the β contribution. Betas were identified by their significant contribution in the low amplitudes. The filtering consisted of removing these low amplitude signals to keep the background noise as low as possible.

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2.3. Morphological and elemental characterisation of the particles by SEM-EDS (Scanning Electron Microscopy - Energy Dispersive Spectroscopy)

A scanning electron microscope (FEI QuantaTM 3D FEG (Field Emission Gun), Eindhoven, The Netherlands) equipped with backscattered and secondary electron detectors was used to determine the size, the morphology and the elemental composition of the particles.

Elemental analyses on particles were realized in-situ via EDS with an Octane Elect Plus detector
(EDAX, Ametek, Tilburg, The Netherlands), which has a surface area of 30 mm². Analyses were
performed with an accelerating voltage of 30 kV to detect U X-ray at 13.6 keV.

- 249 **3. Results and discussion**
- 250

3.1. Theoretical performance of the three methods

251 Theoretical calculations were made to estimate the numbers of (1) α emissions and (2) FT 252 depending on the elemental/isotopic composition and the size of the particle. In this section, we

considered a pure UO₂ particle with a density of 10.97 g·cm⁻³ and a ²³⁵U abundance corresponding to that of fresh FDNPP fuel (²³⁵U = 3.7 atomic %). We considered also spent FDNPP UO₂ fuel of reactor 3 containing 0.7 wt% of Pu [41], MOX fuel of the reactor 3 containing 0.05 g·g⁻¹ of Pu [41], and a natural zircon with a U weight concentration around 1 wt% (determined in zircon crystals from the Saranac Prospect, Bancroft, Ontario) [42].

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3.1.1. Calculation of the number of detected α-tracks

260 The theoretical number of α -tracks detected by α -autoradiography techniques, n_{α} , for a 261 solid sample is given by Eq. (5)

262
$$n_{\alpha} = \frac{\pi}{12} \times K \times N_A \times t_{ini} \times \eta_{\alpha} \times \eta'_{\alpha} \times \rho_P \times d_P^3 \times \left[\frac{C_U}{M_U} \times \left(\sum_{i=234,235,238} (X_{iU} \times \lambda_{iU}) + \sum_{j=1}^n (X_{Udp,j} \times \lambda_{Udp,j})\right) + \frac{C_{Pu}}{M_{Pu}} \sum_{i=239,240} (X_{iPu} \times \lambda_{iPu})\right] (5)$$

264

Where K is the detector efficiency (ratio of the number of registered α tracks and the theoretical α 265 266 emitted), N_A is the Avogadro constant, t_{ini} is the integration time processed for experimental analysis (10 days) (s), η_{α} is the transmission coefficient of α -nuclei within the particle, η'_{α} 267 268 transmission coefficient of α -nuclei between the particle and the detector, ρ_P is the particle density 269 $(g \cdot cm^{-3})$, d_P is the equivalent particle diameter (cm), C_U is the weight concentration of U in the particle (0.88 g·g⁻¹ for pure UO₂, ~10⁻³ g·g⁻¹ for zircon), C_{Pu} is the weight concentration of Pu in 270 the particle (~0.05 g·g⁻¹ for MOX), M_U is the molar mass of U, M_{Pu} is the molar mass of Pu, X_{iU} is 271 the total isotope abundance of 234 U, 235 U and 238 U isotopes in U, X_{iPu} is the total isotope abundance 272 of ²³⁹Pu and ²⁴⁰Pu isotopes in Pu, $X_{Udp,j}$ is the abundance of U decay products: 273 $\sum_{i=1}^{n} (X_{Udp,j} \times \lambda_{Udp,j}) = 6 \times X_{234_{U}} \times \lambda_{234_{U}} = 6 \times X_{238_{U}} \times \lambda_{238_{U}} \text{ in the case of naturally-occurring } U \text{ at secular}$ 274 275 equilibrium or negligible if the U had been chemically purified (U from FDNPP), λ is the

276 disintegration constant of the radionuclides (s⁻¹). According to the literature, the efficiency (K) of 277 IP for α detection is 64% (derived from [27] and [33] based on ²³⁹Pu), while it reaches 51% for the 278 BeaQuant[®] device (derived from [37]).

The determination of the exact value of the product $\eta_{\alpha} \times \eta'_{\alpha}$ was not carried out in this work because it is difficult to estimate. It depends on (1) the particle composition which contained the α -emitting radionuclides, (2) the distribution of the α -emitting radionuclides within the particles (i.e. homogeneously distributed or not, as inclusions, etc.), (3) the radionuclide itself (particle type and emission energy) and (4) the collodion layer thickness into which particles are trapped [37]. Three transmission values were assumed for calculation: 0.01, 0.1 and 1, which correspond to low, medium and high transmissions respectively.

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3.1.2. Calculation of the fission-tracks (FT) detected number

288 The theoretical number of FT (n_f) observed in 2π -steradian can also be estimated for the 289 same types of particles according to Eq. (6) modified equation from [43]:

290
$$n_{FT} = \frac{\pi}{6} \times K_{dry} \times N_A \times t_{ir} \times \eta_{FT} \times \eta'_{FT} \times \rho_P \times$$

291
$$d_P^{3} \times \phi \times \left(\frac{X_{235_U} \times C_U \times \sigma_{235_U}}{M_U} + \frac{X_{239_{Pu}} \times C_{Pu} \times \sigma_{239_{Pu}}}{M_{Pu}}\right)$$
(6)

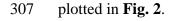
Where K_{dry} is the track registration efficiency of Lexan[®], which is estimated to 96 % ± 4 % [44], N_A is the Avogadro constant, t_{ir} is the irradiation time processed for experimental analysis (s), η_{FT} is the transmission coefficient of fission fragment within the particle , η'_{FT} is the transmission coefficient of fission fragment between the particle and the detector (the product $\eta_{FT} \times \eta'_{FT}$ is determined to be 1), ρ_P is the particle density (g·cm⁻³), d_P is the equivalent particle diameter (cm), ϕ is the thermal neutron flux (1.00 x 10¹⁵ neutrons·cm⁻²) provided by the TRIGA reactor, σ is the

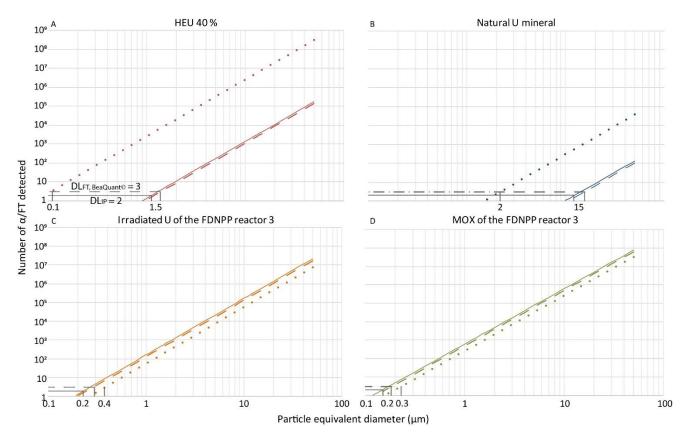
fission cross section under thermal neutrons, $(5.80 \times 10^{-22} \text{ cm}^2 \text{ for } {}^{235}\text{U} \text{ and } 7.42 \times 10^{-22} \text{ cm}^2 \text{ for}$ 298 239 Pu), X_{235U} is the total isotope abundance of 235 U in U, X_{239Pu} is the total isotope abundance of 299 ²³⁹Pu in Pu, C_U is the weight concentration of U in the particle (0.88 g·g⁻¹ for pure UO₂, ~10⁻³ g·g⁻¹ 300 ¹ for zircon), C_{Pu} is the weight concentration of Pu in the particle (~0.05 g·g⁻¹ for MOX), M_U is the 301 302 molar mass of U, M_{Pu} is the molar mass of Pu.

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3.1.3. Comparison of the α -autoradiography and FT methods theoretical 305 performance for the actinide-bearing particles of interest

306 These calculations are applied on elemental compositions chosen previously. Results are





308

Fig. 2 Graphical representation of theoretical calculation results for $\eta_{\alpha} \times \eta'_{\alpha} = 1$ (high transmission) 309 310 expressed as the number of detected α-tracks/FT plotted against particle size, particle composition

and technique used, with logarithmic scales for both axes. A: HEU 40 %; B: Natural U mineral; C: Irradiated U from the FDNPP reactor 3; D: MOX of the FDNPP reactor 3 (for graphs C and D the burnup average values determined by [41] are considered). Colored solid lines represent results for IP; colored dashed line represent results for BeaQuant[®]; colored dotted lines represent results for FT. The black lines represent the detection limits of each method in our analytical conditions (see **Table 1**). Results for $\eta_{\alpha} \times \eta'_{\alpha} = 0.1$ (medium transmission) and $\eta_{\alpha} \times \eta'_{\alpha} = 0.01$ (low transmission) are given in supplementary information B.

318

319 The theoretical calculations demonstrate the impact of the transmission coefficient and 320 particle size and composition on detecting of α -emissions. For a transmission coefficient of the α nuclei equal to 1 ($\eta_{\alpha} \times \eta'_{\alpha} = 1$), the theoretical calculations show that only HEU particles larger 321 than 1 µm can be detected by IP and the BeaQuant[®], whereas HEU particles larger than 0.1 µm 322 323 can theoretically be detected with the FT method. Regarding the naturally occurring U particles, 324 they can be detected by all of three methods respectively with a full transmission of the α -nuclei for equivalent diameters larger than 15 µm for IP and BeaQuant[®], and larger than 2 µm for FT. By 325 326 contrast, detection of irradiated U particles and reactor 3 MOX particles with equivalent diameters larger than 0.2 μ m is possible even with the lowest transmission factor ($\eta_{\alpha} \times \eta'_{\alpha} = 0.01$) by means 327 328 of the three tested methods. Actually, as explained by [45], a small fraction of Pu in U (in the 10⁻³ -10^{-2} range) strongly increases the number of α -emissions. 329

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331

Accordingly, the FT method may be the method offering the best chance of detecting U bearing-particles from FDNPP with an equivalent diameter greater than 0.3 µm.

332

333 3.2. Experimental results

334 Results of the comparison of the three localisation methods for particles that create FT or 335 α -track clusters above the detection limits are summarized in **Table 1**. Regarding the BeaQuant[®], 336 we gathered the results obtained with the two-acquisition methods (methods 1 and 2), as the same 337 post-processing was implemented for the two methods.

338

339	Table 1 Results obtained with the three-localisation methods tested in this study. "/" corresponds
340	to sample not analysed. For IP and BeaQuant® values of background noise and detection limits
341	reported here are the averages for all of our experiments for all samples. Background noise and
342	detection limits for the FT, IP and BeaQuant® methods are expressed in number of fission tracks
343	observed per cluster, number of α -tracks observed per cluster, and counts per pixel, respectively.

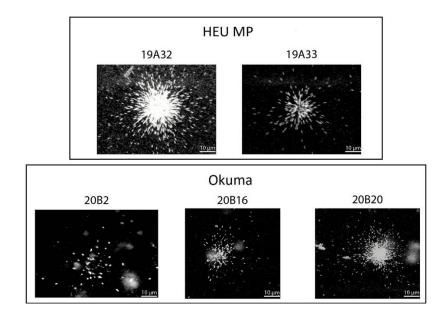
		Detection	Number of clusters detected for samples					
Localisation method	Background noise	limit (DL) for a	HEU MP	HEU MP	Okuma	Okuma	Okuma	Control
incense	noise	cluster	19A32	19A33	20B2	20B16	20B20	Control
FT	1	3	4	4	3	2	1	/
IP	2	6	0	/	9	1	14	/
BeaQuant®	1	3	/	0	0	3	1	1

344

Background noise is determined for parts of the detector without radioactive particles. Thedetection limit values are calculated according to the Eq. (7):

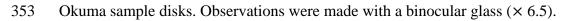
$$DL = mean_{background} + 3 \times RSD \tag{7}$$

The FT method allows locating particles that contain fissile nuclei for both HEU MP and Okuma disks (**Fig. 3**). The advantage of this method that it is insensitive to the background associated with the high activity of ¹³⁷Cs.



351

Fig. 3 Images of large FT clusters (more than 30 tracks) observed for HEU MP sample disks and



- 354
- 355 The IP method revealed pixel clusters with significant PSL above DL only for Okuma disks
- 356 (**Fig. 4.c**).

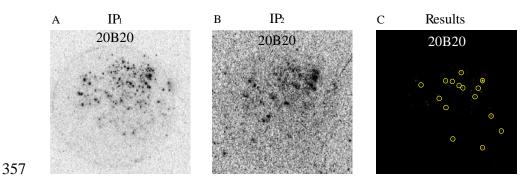


Fig. 4 A: Autoradiography image of IP1 without correction; B: Autoradiography image of IP2
without correction; C: Autoradiography image after processing which allowed localising
significant clusters of pixels. Yellow dots correspond to clusters of pixels above DL.

The BeaQuant[®] has a homogeneous background whatever the sample. Clusters are observed for the Okuma sample (disks 20B16, and 20B20) and for the Control sample (IAEA-472). The detection of one pixel above background for the Control sample, which has a relatively low ¹³⁷Cs activity (Supplementary information C), suggests that the influence of the β - γ activity is not fully removed. The background observed could be link to radon levels induce by long exposure time and Lexan[®] disks. However, this method did not allow the detection of α -emitting particles for the HEU MP and for the disk 20B2 of the Okuma.

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370

371

radioactive particles in a sediment sample from FDNPP

3.3. Comparison of methods for the localisation of pure U-bearing particles and

372 Here we compared the performance of the methods applied to the HEU MP samples that 373 contain only pure micrometric U particles. For these samples, only the FT method allowed the 374 detection of particles as shown in **Table 1**. The inefficiency of IP and BeaQuant[®] methods may be 375 associated with the particle diameter correlated to the transmission coefficient. As demonstrated 376 in Fig. 2, for transmission coefficients respectively equal to 1, 0.1, and 0.01, the equivalent 377 diameter of particles should be higher than 1, 2, and 5 µm to be detected. Accordingly, for detecting pure HEU particles (²³⁵U atomic abundance of 40%), only the FT method was efficient. Indeed, 378 379 SIMS measurements show that most of the HEU particles have diameters between 0.25 to 0.4 μ m. 380 This can be explained by (1) the fact that particles are too small (in the micron range) and thus do 381 not contain enough α -emitting nuclei, (2) the transmission coefficient or yield for IP and 382 BeaQuant[®] methods are too low, or (3) a fading effect may limit their detection by IP [46]. 383 Consequently, α activity could not be detected within 10 days.

The three methods were also applied to the sediment sample Okuma collected near the FDNPP. As explained previously, aliquots of the sample were deposited on three Lexan[®] disks. The FT method enabled the detection of three particles in the disk 20B2, two particles in the disk 20B16 and one in disk 20B20.

However, the IP method did not allow the detection of α-emitting particles at the same locations on the FT disks, although several clusters of pixels with an intensity higher than the detection limit were detected at other locations (**Fig. 5**). The pixel clusters (set of pixels with a spatial resolution of 150 dpi (169 μ m)) can be induced by β-γ emissions of radio-cesium and by radon levels which may impact background levels with long exposure time, that were not fully removed by the data treatment. It cannot be mineral particles that contain naturally occurring U as these particles would also have been detected by means of the FT method.

395

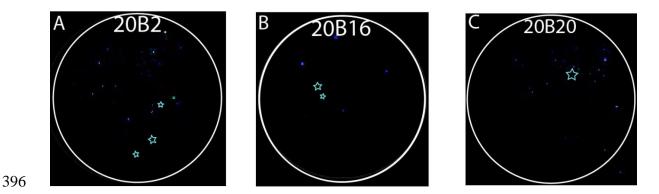


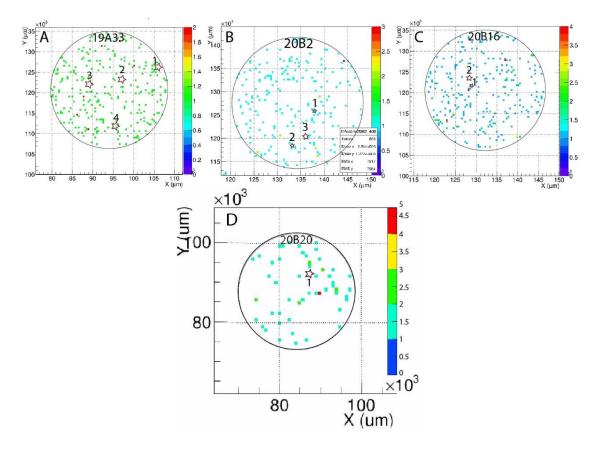
Fig. 5 Comparison between FT and IP results for disks 20B20 (A), 20B16 (B), and 20B2 (C). Star
shapes correspond to the location of FT clusters and blue points correspond to clusters of pixels
detected by IP (reading with a resolution of 150 dpi (169 µm)).

400

401 The BeaQuant[®] device showed few pixels with values above the DL, i.e. with 3 or 4 counts
402 (pixel resolution: 400 μm). However, again, when results are compared to those of the FT method,

403 no correlation is observed whereas any types of α -emitter-bearing particles would have been 404 detected with the FT method. So false pixel detection on the Control sample and blank samples 405 may suggest that the BeaQuant[®] settings and/or data treatment does not fully remove β 406 contributions (**Fig. 6**). Actually, considering the higher ¹³⁷Cs activity of the Fukushima samples 407 compared to that of Control sample, we cannot exclude the possibility of an increase of the 408 background due to the β emissions of radio-cesium as observed with IP.

409 To summarise, no correspondence was observed between on one side the pixels above the 410 DL recorded with the BeaQuant and the clusters of α -tracks above DL with the IP method and on 411 the other side the localisation of FT clusters that exceed the DL.



413 Fig. 6 Comparison between FT and BeaQuant[®] results (400 × 400 μm²). A: HEU MP sample; B:
414 Okuma disk 20B2; C: Okuma disk 20B16; D: Okuma disk 20B20. Purple star shapes correspond

415 to the location of FT clusters and coloured pixels correspond to counts detected with the
416 BeaQuant[®].

417

418 Combination of theoretical calculation and experimental comparison of the three 419 localisation methods revealed that only the FT method allows locating U-bearing particles for the 420 Okuma and HEU MP samples. Several hypotheses can be put forward to explain the absence of 421 particle detection by the two autoradiography methods:

- 422 (1) Particles were too small, and consequently α -activities were too low to induce 423 detectable α activity per surface unit (IP pixel size 169 × 169 μ m²; BeaQuant[®] pixel 424 size 400 × 400 μ m²).
- 425 (2) The layer of organic polymer (collodion) used to fix the sediment particles may 426 significantly reduce the number of α emerging from the surface of the sample disks 427 (low η'_{α}). It should be noted that the collodion thickness is not controlled or known.
- 428 (3) The self-absorption phenomenon may be too high if the U is embedded in the form of 429 nanoparticles within larger radioactive microparticles mainly made of silica for 430 instance (low η_{α}). Actually, according to the literature, α -emissions from actinides have 431 a penetration of only a few μ m in U-rich minerals [47].
- 432 (4) A fading effect may occur for the IP method, which can lead to loss of information for
 433 low activities with the two-week long exposure time of this study [46]. Nevertheless,
 434 such a long exposure time was mandatory to increase the probability of detecting
 435 micrometric particles.

The size and U concentration of the three particles that gave large clusters of FT in disks
20B2, 20B16 and 20B20 were evaluated by SEM-EDS. The particle of the disk 20B2 (Fig. 7a1)

has a pseudo-spherical shape with a diameter of ~40 μ m. The particle on the disk 20B16 (**Fig. 7a**₂) has a rod shape with a length of ~50 μ m in the longest dimension. The particle on the disk 20B20 (**Fig. 7a**₃) has no specific shape with a length of ~40 μ m. These particles contain elements that are abundant in many minerals including Si, O, Al and Fe (**Fig. 7b**₁, **b**₂ **and b**₃) although they also include the following elements: Ti, Mg, Na, K, P and S for disk 20B2, and Zr for disks 20B16 and 20B20.

444

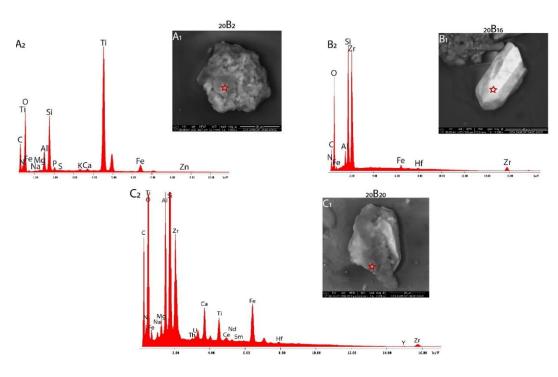


Fig. 7 SEM images and EDS spectra of the three particles that produces large clusters (A₁, A₂, B₁, B₂, C₁ and C₂) are presented. A₁: SEM image of the particle of the disk 20B2; A₂: EDS spectrum of the particle of the 20B2 disk; B₁: SEM image of the zircon of disk 20B16; B₂: EDS spectrum of zircon of the disk 20B16; C₁: SEM image of the zircon of the disk 20B20; C₂: EDS spectrum of zircon of the disk 20B16; C₁: SEM image of the zircon of the disk 20B20; C₂: EDS spectrum of zircon of the disk 20B20. Red stars represent the spots where EDS spectra were performed.

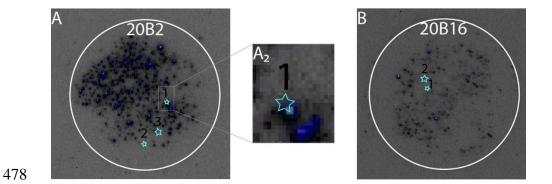
Given their elemental composition and morphology, it is assumed that the particles detected in disks 20B16 and 20B20 are zircons. In contrast, the elemental composition of the particle observed on disk 20B2 is very close to that of the FDNPP-originating particles observed by [4, 10]). However, U was not detected by EDS in the particle found in disk 20B2 despite the observation of a large FT cluster, which may indicate that its concentration remains below the EDS detection limit, i.e. less than 1 wt%.

458 In the current analytical conditions and in the investigated samples, among the three 459 methods tested, only the FT method allows localising U-bearing particles (natural or 460 anthropogenic). However, as shown by SEM-EDS results on disks 20B16 and 20B2, the FT 461 method itself does not distinguish between natural U particles and those that may originate from 462 FDNPP. Coupling of FT imaging and IP β - γ imaging could allow differentiating between 463 naturally-occurring U in minerals and anthropogenic U-bearing particles from FDNPP. Indeed, the 464 correlation of a FT cluster with an IP pixel cluster of radio-cesium β - γ signals would strongly 465 suggest that the corresponding particle originates from FDNPP. This method was implemented on 466 these three disks. Instead of the IP data treatment described in Eq. 4, the IP1 autoradiography image was treated in order to detect large clusters induced by β - γ emission. The obtained β - γ 467 468 images were then superimposed with the FT imaging.

In the case of the zircon particles, no correspondence with β signals was found (**Fig. 8B**), whereas in the case of the potential anthropogenic particle a coincidence was clearly observed (**Fig. 8A**). Therefore, this particle has great chances to be anthropogenic, released by FDNPP. Further investigations will be performed on that particle in order to identify its origin. These preliminary results suggest that coupling of FT and β - γ IP can be an alternative technique to determine the source of U-bearing particles. Such result could be improved by coupling FT and

475 BeaQuant[®] in β mode thanks to its better pixel resolution and less sensitivity to the ambient 476 background.

477



479Fig. 8 Superimposition of the localisation FT and β-γ cluster respectively for disks 20B2 and48020B16 (A and B). Light blue stars represent FT clusters.

481

482 **4.** Conclusions

483 The analytical and experimental results of this study demonstrate that under the chosen 484 analytical conditions and the selected samples the fission track method is more effective than the 485 two tested α -autoradiography methods to detect and localise U-bearing particles.

In this study, HEU MP, as well as particles containing fissile nuclei (natural, or anthropogenic U) in a sediment sample collected in the vicinity of the FDNPP were only recorded by the fission track method. However, the current study also demonstrated that, for a sediment sample, this method alone does not allow to distinguish (1) particles released by FDNPP which contain fragments of nuclear fuel from (2) mineral particles which contain naturally-occurring U. Indeed, our experiments showed that natural zircon minerals with a diameter larger than 10 μ m were also detected by means of the FT method. On the contrary, the two α -radiography methods

493 did not provide any results for our samples. This may probably be explained by a too high auto-494 absorption (1) either in the collodion layer, which is nevertheless required to prevent 495 contamination of the detector by radioactive particles, or (2) possibly within the particles themselves. Nevertheless the results obtained for the two a-radiography methods confirm that 496 497 presently the developed data treatment did not enable full removal of the contribution of the β 498 and/or γ emitters on the α images. Thus, the implementation of a method coupling the FT efficiency 499 for the detection of α -emitters and the IP efficiency for the detection of β - γ -emitters was tested in 500 order to attempt the discrimination between naturally occurring U-bearing particles and 501 anthropogenic particles. Results are promising because they enabled the distinction between zircon 502 particles and a particle that could potentially come from FDNPP. Restricting the analysis to 503 particles detected by both methods will definitely avoid isolating U-rich minerals and significantly 504 increase the probability of identifying anthropogenic radioactive particles.

To conclude, this study opens new perspective for the detection and localisation of Ubearing particles in sediment/soil samples. The combined methodology will be applied to more sediment and soil samples collected in areas located within the main radioactive contamination plume of FDNPP in order to isolate U-bearing particles. The methodology can also probably be improved by coupling the FT method with the BeaQuant® instrument because of its better sensitivity and spatial resolution than IP.

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520 **References**

- 521 1. Kashparov VA (2003) Hot Particles at Chernobyl. 10
- Salbu B, Krekling T, Oughton DH, et al (1994) Hot particles in accidental releases from
 Chernobyl and Windscale nuclear installations. Analyst 119:125–130.
 https://doi.org/10.1039/AN9941900125
- Sandalls FJ, Segal MG, Victorova N (1993) Hot particles from Chernobyl: A review. Journal of Environmental Radioactivity 18:5–22. https://doi.org/10.1016/0265-931X(93)90063-D

Abe Y, Iizawa Y, Terada Y, et al (2014) Detection of Uranium and Chemical State Analysis
of Individual Radioactive Microparticles Emitted from the Fukushima Nuclear Accident
Using Multiple Synchrotron Radiation X-ray Analyses. Anal Chem 86:8521–8525.
https://doi.org/10.1021/ac501998d

5. Adachi K, Kajino M, Zaizen Y, Igarashi Y (2013) Emission of spherical cesium-bearing
particles from an early stage of the Fukushima nuclear accident. Scientific reports 3:2554.
https://doi.org/10.1038/srep02554

534 6. Imoto J, Ochiai A, Furuki G, et al (2017) Isotopic signature and nano-texture of cesium-rich
535 micro-particles: Release of uranium and fission products from the Fukushima Daiichi
536 Nuclear Power Plant. Scientific Reports 7:5409. https://doi.org/10.1038/s41598-017-05910537 z

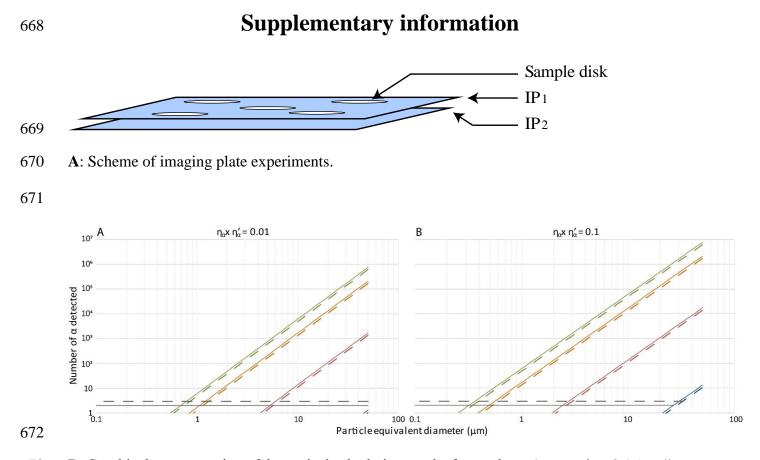
- Furuki G, Imoto J, Ochiai A, et al (2017) Caesium-rich micro-particles: A window into the
 meltdown events at the Fukushima Daiichi Nuclear Power Plant. Scientific Reports 7:1–10.
 https://doi.org/10.1038/srep42731
- Kurihara Y, Takahata N, Yokoyama TD, et al (2020) Isotopic ratios of uranium and caesium in spherical radioactive caesium-bearing microparticles derived from the Fukushima Dai-ichi Nuclear Power Plant. Scientific Reports 10:1–10. https://doi.org/10.1038/s41598-020-544 59933-0

- 545 9. Kurihara E, Takehara M, Suetake M, et al (2020) Particulate plutonium released from the
 546 Fukushima Daiichi meltdowns. Science of The Total Environment 743:140539.
 547 https://doi.org/10.1016/j.scitotenv.2020.140539
- Martin PG, Griffiths I, Jones CP, et al (2016) In-situ removal and characterisation of uraniumcontaining particles from sediments surrounding the Fukushima Daiichi Nuclear Power Plant.
 Spectrochimica Acta Part B: Atomic Spectroscopy 117:1–7. https://doi.org/10.1016/j.sab.2015.12.010
- Martin PG, Louvel M, Cipiccia S, et al (2019) Provenance of uranium particulate contained
 within Fukushima Daiichi Nuclear Power Plant Unit 1 ejecta material. Nature
 Communications 10:1–7. https://doi.org/10.1038/s41467-019-10937-z
- Martin PG, Jones CP, Cipiccia S, et al (2020) Compositional and structural analysis of
 Fukushima-derived particulates using high-resolution x-ray imaging and synchrotron
 characterisation techniques. Scientific Reports 10:1636. https://doi.org/10.1038/s41598-020 58545-y
- 13. Ochiai A, Imoto J, Suetake M, et al (2018) Uranium Dioxides and Debris Fragments Released
 to the Environment with Cesium-Rich Microparticles from the Fukushima Daiichi Nuclear
 Power Plant. Environ Sci Technol 52:2586–2594. https://doi.org/10.1021/acs.est.7b06309
- 562 14. Satou Y, Sueki K, Sasa K, et al (2018) Analysis of two forms of radioactive particles emitted
 563 during the early stages of the Fukushima Dai-ichi Nuclear Power Station accident.
 564 Geochemical Journal 52:137–143. https://doi.org/10.2343/geochemj.2.0514
- 565 15. Satou Y, Sueki K, Sasa K, et al (2016) First successful isolation of radioactive particles from
 566 soil near the Fukushima Daiichi Nuclear Power Plant. Anthropocene 14:71–76.
 567 https://doi.org/10.1016/j.ancene.2016.05.001
- Igarashi Y, Kogure T, Kurihara Y, et al (2019) A review of Cs-bearing microparticles in the
 environment emitted by the Fukushima Dai-ichi Nuclear Power Plant accident. Journal of
 Environmental Radioactivity 205–206:101–118.
 https://doi.org/10.1016/j.jenvrad.2019.04.011
- 572 17. Aarkrog A (1971) Radioecological Investigations of Plutonium in an Arctic Marine
 573 Environment. Health Physics 20:31–47
- 574 18. Eriksson M, Ljunggren K, Hindorf C (2002) Plutonium hot particle separation techniques
 575 using real-time digital image systems. Nuclear Instruments and Methods in Physics Research
 576 Section A: Accelerators, Spectrometers, Detectors and Associated Equipment 488:375–380.
 577 https://doi.org/10.1016/S0168-9002(02)00438-2
- 578 19. Jiménez-Ramos MC, García-López J, García-Tenorio R, García-León M (2009)
 579 Characterization of terrestrial hot particles from the Palomares accident using destructive and
 580 non-destructive analytical techniques. Radioprotection 44:345–350.
 581 https://doi.org/10.1051/radiopro/20095067

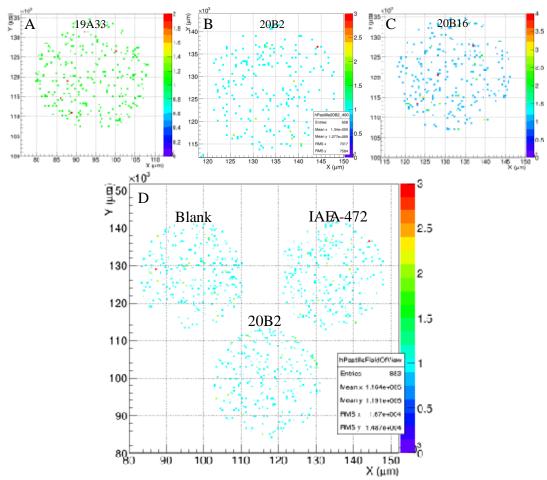
- 582 20. Jiménez-Ramos MC, García-Tenorio R, Vioque I, et al (2006) Presence of plutonium
 583 contamination in soils from Palomares (Spain). Environmental Pollution 142:487–492.
 584 https://doi.org/10.1016/j.envpol.2005.10.030
- 585 21. Pöllänen R, Ketterer ME, Lehto S, et al (2006) Multi-technique characterization of a nuclearbomb particle from the Palomares accident. Journal of Environmental Radioactivity 90:15–28. https://doi.org/10.1016/j.jenvrad.2006.06.007
- 588 22. Donohue DL (2002) Peer Reviewed: Strengthened Nuclear Safeguards. Anal Chem 74:28 A589 35 A. https://doi.org/10.1021/ac021909y
- Jaegler H, Pointurier F, Onda Y, et al (2018) Plutonium isotopic signatures in soils and their
 variation (2011-2014) in sediment transiting a coastal river in the Fukushima Prefecture,
 Japan. Environmental Pollution 240:167–176. https://doi.org/10.1016/j.envpol.2018.04.094
- Esaka KT, Esaka F, Inagawa J, et al (2004) Application of Fission Track Technique for the
 Analysis of Individual Particles Containing Uranium in Safeguard Swipe Samples. Jpn J Appl
 Phys 43:L915. https://doi.org/10.1143/JJAP.43.L915
- Sardini P, Angileri A, Descostes M, et al (2016) Quantitative autoradiography of alpha particle emission in geo-materials using the BeaverTM system. Nuclear Instruments and Methods in Physics Research Section A: Accelerators, Spectrometers, Detectors and Associated Equipment 833:15–22. https://doi.org/10.1016/j.nima.2016.07.003
- Furuki G, Imoto J, Ochiai A, et al (2017) Caesium-rich micro-particles: A window into the
 meltdown events at the Fukushima Daiichi Nuclear Power Plant. Scientific Reports 7:1–10.
 https://doi.org/10.1038/srep42731
- 403 27. Haudebourg R, Fichet P (2016) A non-destructive and on-site digital autoradiography-based
 404 tool to identify contaminating radionuclide in nuclear wastes and facilities to be dismantled.
 405 J Radioanal Nucl Chem 309:551–561. https://doi.org/10.1007/s10967-015-4610-7
- Chartin C, Evrard O, Onda Y, et al (2013) Tracking the early dispersion of contaminated
 sediment along rivers draining the Fukushima radioactive pollution plume. Anthropocene
 1:23–34. https://doi.org/10.1016/j.ancene.2013.07.001
- 609 29. IAEA (2019) Environmental Sample Particle Analysis Interlaboratory Comparison
- 30. Bonnet T, Comet M, Denis-Petit D, et al (2013) Response functions of imaging plates to
 photons, electrons and 4He particles. Review of Scientific Instruments 84:103510.
 https://doi.org/10.1063/1.4826084
- 613 31. Chen B, Zhuo W, Kong Y (2011) Identification and counting of alpha tracks by using an
 614 imaging plate. Radiation Measurements 46:371–374.
 615 https://doi.org/10.1016/j.radmeas.2011.01.002

- Billon S, Sardini P, Leblond S, et al (2018) MAUD PROJECT-Comparative study of 3
 Digital Autoradiography Techniques for characterization: Gas Detector -Solid Scintillation
 Detector -Phosphor Screens
- 619 33. Rahman NM, Iida T, Yamazawa H, Moriizumi J (2006) Determination of Alpha Particle
 620 Detection Efficiency of an Imaging Plate (IP) Detector. Jpn J Health Phys 41:272–278.
 621 https://doi.org/10.5453/jhps.41.272
- 622 34. Fichet P, Bresson F, Leskinen A, et al (2012) Tritium analysis in building dismantling process
 623 using digital autoradiography. J Radioanal Nucl Chem 291:869–875.
 624 https://doi.org/10.1007/s10967-011-1423-1
- 35. Radulović V, Kolšek A, Fauré A-L, et al (2018) Qualification of heavy water based
 irradiation device in the JSI TRIGA reactor for irradiations of FT-TIMS samples for nuclear
 safeguards. Nuclear Instruments and Methods in Physics Research Section A: Accelerators,
 Spectrometers, Detectors and Associated Equipment 885:139–144.
 https://doi.org/10.1016/j.nima.2017.12.046
- 630 36. Donnard J, Berny R, Carduner H, et al (2009) The micro-pattern gas detector PIM: A multi631 modality solution for novel investigations in functional imaging. Nuclear Instruments and
 632 Methods in Physics Research Section A: Accelerators, Spectrometers, Detectors and
 633 Associated Equipment 610:158–160. https://doi.org/10.1016/j.nima.2009.05.186
- 634 37. Billon S, Sardini P, Angileri A, et al (2020) Quantitative imaging of 226Ra ultratrace
 635 distribution using digital autoradiography: Case of doped celestines. Journal of
 636 Environmental Radioactivity 217:106211. https://doi.org/10.1016/j.jenvrad.2020.106211
- 38. Donnard J, Arlicot N, Berny R, et al (2009) Advancements of labelled radio-pharmaceutics
 imaging with the PIM-MPGD. J Inst 4:P11022–P11022. https://doi.org/10.1088/17480221/4/11/P11022
- 640 39. Donnard J, Thers D, Servagent N, Luquin L (2009) High Spatial Resolution in β-Imaging
 641 With a PIM Device. IEEE Transactions on Nuclear Science 56:197–200.
 642 https://doi.org/10.1109/TNS.2008.2005673
- 40. Angileri A, Sardini P, Beaufort D, et al (2020) Mobility of daughter elements of 238U decay chain during leaching by In Situ Recovery (ISR): New insights from digital autoradiography.
 Journal of Environmental Radioactivity 220–221:106274.
 https://doi.org/10.1016/j.jenvrad.2020.106274
- 647 41. Nishihara K, Iwamoto H, Suyama K (2012) Estimation of fuel compositions in Fukushima648 Daiichi nuclear power plant. 202. https://doi.org/JAEA-Data/Code-2012-018
- 42. Nasdala L, Hanchar JM, Rhede D, et al (2010) Retention of uranium in complexly altered
 zircon: An example from Bancroft, Ontario. Chemical Geology 269:290–300.
 https://doi.org/10.1016/j.chemgeo.2009.10.004

- Kalsi PC, Sawant PD, Ramaswami A, Manchanda VK (2007) Track etching characteristics
 of polyester track detector and its application to uranium estimation in seawater samples. J
 Radioanal Nucl Chem 273:473–477. https://doi.org/10.1007/s10967-007-6845-4
- 44. Khan HA, Durrani SA (1972) Efficiency calibration of solid state nuclear track detectors.
 Nuclear Instruments and Methods 98:229–236. https://doi.org/10.1016/0029-554X(72)90103-6
- 45. Jaegler H, Pointurier F, Onda Y, et al (2019) Method for detecting and characterising
 actinide-bearing micro-particles in soils and sediment of the Fukushima Prefecture, Japan. J
 Radioanal Nucl Chem 321:57–69. https://doi.org/10.1007/s10967-019-06575-w
- 46. Muuri E, Sorokina T, Donnard J, et al (2019) Electronic autoradiography of 133Ba particle
 emissions; diffusion profiles in granitic rocks. Applied Radiation and Isotopes 149:108–113.
 https://doi.org/10.1016/j.apradiso.2019.04.026
- 47. Angileri A, Sardini P, Donnard J, et al (2018) Mapping 238U decay chain equilibrium state
 in thin sections of geo-materials by digital autoradiography and microprobe analysis. Applied
 Radiation and Isotopes 140:228–237. https://doi.org/10.1016/j.apradiso.2018.06.018



B: Graphical representation of theoretical calculation results for products A. $\eta_{\alpha} \ge \eta_{\alpha} \ge \eta_{$



681 C: BeaQuant® results after the application of low amplitude filter A. Quality control disk 19A33;

Fukushima samples; B. 20B2 (two acquisitions) and C. 20B16; D. Blank and IAEA-472 samples.