

## Improving the quantification of sediment source contributions using different mathematical models and spectral preprocessing techniques for individual or combined spectra of ultraviolet-visible, near-and middle-infrared spectroscopy

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#### 24 Abstract

In recent years, several sediment fingerprinting studies have used ultraviolet-visible (UV-25 26 Vis), near-infrared (NIR) and middle-infrared (MIR) spectroscopy as a low cost, non-27 destructive and fast alternative to obtain tracer properties to estimate sediment source 28 contributions. For this purpose, partial last square regression (PLSR) has often been used to 29 build predictive parametric models. However, spectra preprocessing and more robust and 30 non-parametric models such as support vector machines (SVM) has gained little attention 31 in these studies. Accordingly, the objectives of the current research were to evaluate (i) the 32 accuracy of two multivariate methods (PLSR and SVM), (ii) the effect of eight spectra 33 preprocessing techniques, and (iii) the effect of using the information contained in the UV-34 Vis, NIR and MIR regions considered either separately or in combination on sediment 35 source apportionment. The estimated source contribution was then compared with 36 contributions obtained by the conventional fingerprinting approach based on geochemical 37 tracers. This study was carried out in the Arvorezinha catchment (1.23 km<sup>2</sup>) located in 38 southern Brazil. Forty soil samples were collected in three main potential source (cropland surface, unpaved roads and stream channels) and twenty-nine suspended sediment 39 40 samples collected at the catchment outlet during nine rainfall-runoff events were used in this study. Both PLSR and SVM models showed a higher accuracy when calibrated and 41 42 validated with the spectra submitted to spectral processing when compared to the direct 43 use of the raw spectra. The best model results were obtained with PLSR and SVM 44 mathematical models associated with the spectral preprocessing techniques 1st derivative 45 Savitzky-Golay (SGD1), normalization (NOR) and combining NOR+SGD1 in the UV-Vis+NIR+MIR. The lowest errors were observed when the UV-Vis+NIR+MIR bands were 46 47 combined due to the gain in information and, consequently, the increase in discriminant power achieved by the models. Despite the good accuracy of the models calibrated and 48 validated with the mixtures, significant errors remain when results of source contributions 49 are compared to those obtained with the conventional sediment fingerprinting technique 50 based on geochemical tracers. Nevertheless, the magnitude of the contributions calculated 51 by the spectroscopy and geochemical approaches remains very similar for all sources, 52 especially when using the SVM-UV-Vis+NIR+MIR model. Therefore, spectroscopy proved to 53 be a fast, cheap and accurate technique, offering an alternative to the conventional 54 geochemical approach for discriminating sediment source contributions in agricultural 55 56 catchments located in subtropical regions.

57 Keywords: soil erosion, alternative fingerprints, machine learning, Vis-NIR-MIR database,58 spectral preprocessing.

The sustainable production of food, fiber and biofuel remains limited by soil erosion 60 (Comino et al., 2015; Erkossa et al., 2015; Seutloali and Beckedahl, 2015; Taguas et al., 61 62 2015). The inadequate management of soils and the lack of runoff control exposes the soil to erosive agents, accelerating the processes of mobilization and transfer of sediments to 63 the drainage network (Minella et al., 2014), along with the transportation of contaminants 64 like pesticides (Magnusson et al., 2013; Yahia and Elsharkawy, 2014) and phosphorus 65 (Dodd et al., 2014; Dodd and Sharpley, 2015; Poulenard et al., 2008). Erosion control and 66 conservation of soil hydrological functionalities are essential to meet the demand in food 67 68 production as well as to maintain the quality of water resources (Didoné et al., 2015; Merten et al., 2015). To better understand the occurrence of erosion processes at the 69 70 catchment scale and to mitigate the problems arising from river overflow and excessive 71 sediment production, it is first necessary to have quantitative information on the sources 72 delivering sediment to the river systems.

To this end, the use of the sediment fingerprinting approach quantifies the contribution of 73 non-point based sediment sources through the use of a variety of tracers, including 74 75 geochemical properties and radionuclides (D'Haen et al., 2012; Davis and Fox, 2009; 76 Haddadchi et al., 2013; Koiter et al., 2013; Walling and Woodward, 1995). However, the large-scale application of the sediment fingerprinting technique requires a large number of 77 78 chemical analyses, which have a significant cost and which are relatively time consuming. 79 However, the use of other soil and sediment characteristics may complement or provide a powerful alternative information in source identification, such as reflectance spectroscopy 80 81 (Cooper et al., 2014).

82 Reflectance spectroscopy in the visible and infrared ranges provides an efficient, low-cost, fast and non-destructive method easily applicable to soil and sediment samples to quantify 83 various physical, chemical and biological properties (McBratney et al., 2006; Viscarra Rossel 84 et al., 2006). Poulenard et al. (2009) were pioneers in the use of diffuse infrared 85 spectroscopy (Diffuse Reflectance Infrared Fourier Transform Spectroscopy - DRIFTS) to 86 87 trace the origin of sediments in river catchments. The method was successfully used to 88 discriminate and predict the respective contribution of surface and subsurface sources to sediment, as well as to discriminate soils developed on different lithology (Poulenard et al., 89 90 2012). Since then, several studies have been developed using the spectroscopic method to trace sediment sources (Table 1). Only 28 scientific articles had been published around the 91 92 world after that of Poulenard et al., (2009) by early 2020, representing an average of about 93 2.5 scientific publications per year (Fig. 1). They include studies in the ultraviolet-visible (UV-Vis) and near-infrared (NIR) ranges in the French Alps (Legout et al., 2013) and the 94 Southern France (Uber et al., 2019), Luxemburg (Martínez-Carreras et al., 2010c, 2010b, 95 96 2010a), Spain (Brosinsky et al., 2014a, 2014b), Ethiopia (Verheyen et al., 2014), South Africa 97 (Pulley et al., 2018; Pulley and Rowntree, 2016), Argentina (Batistelli et al., 2018), the 98 United Kingdom (Collins et al., 2014), Canada (Barthod et al., 2015; Boudreault et al., 2018; Liu et al., 2017), Brazil (Tiecher et al., 2016, 2015; Valente et al., 2020) and Iran (Nosrati et 99 100 al., 2020). In the middle infrared region (MIR), sediment fingerprinting studies were carried 101 out in France (Poulenard et al., 2012, 2009), Mexico (Evrard et al., 2013), the United 102 Kingdom (Vercruysse and Grabowski, 2018), China (Liu et al., 2019), Brazil (Tiecher et al., 103 2017) and in a transnational river catchment covering part of Switzerland, France and 104 Germany (Chapkanski et al., 2019). Many of these studies have shown a good agreement 105 between the results obtained with the spectroscopic method and those provided by the

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conventional approach based on geochemical and / or radionuclide properties (Evrard et
al., 2013; Legout et al., 2013; Martínez-Carreras et al., 2010c; Tiecher et al., 2015, 2016,
2017; Verheyen et al., 2014).

109 In addition to the more qualitative attempts that combine the use of discriminant analysis with spectroscopy (Chapkanski et al., 2019), previous studies conducted to trace sediment 110 111 source contributions using spectroscopy (Table 1) can be divided into three main groups, 112 according to the way they used the spectroscopic information. The first group uses color 113 parameters extracted from the visible range (Barthod et al., 2015; Martínez-Carreras et al., 2010c, 2010b, 2010a; Pulley et al., 2018; Pulley and Rowntree, 2016) and other spectral 114 115 characteristics (spectral features and overtones) (Brosinsky et al., 2014b, 2014a; Collins et 116 al., 2013, 2014) in an optimized mixed linear model, separately or in combination with 117 conventional geochemical tracers (Tiecher et al., 2015). The second group of studies uses spectroscopic information to generate mathematical models using the least squares 118 method (PLSR) to estimate the concentrations of geochemical tracers, which in turn are 119 120 introduced in a mixed linear model optimized to estimate the contribution of sediment 121 sources (Vis-NIR-SWIR - Martínez-Carreras et al., 2010b).

The third group of studies directly uses the entire spectrum (Batistelli et al., 2018; Evrard et al., 2013; Poulenard et al., 2012, 2009; Tiecher et al., 2017, 2016, 2015; Verheyen et al., 2014) or the spectrum combined with color parameters extracted from the visible range (Legout et al., 2013). They estimate the source contributions in sediment samples after generating a model (*Partial Last Squares Regression* – PLSR) calibrated using artificial mixtures combining potential sediment sources in variable proportions. To date, no study has been conducted combining the bands of UV-Vis, NIR and / or MIR, although it may be

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expected that predictions of soil properties may be improved using the combined range of these regions of the electromagnetic spectrum (Knox et al., 2015; Reeves, 2010; Soriano-Disla et al., 2014) due to the sum of distinct information that is added to the models (Viscarra Rossel et al., 2006).

This third approach has the advantage of using all the spectral information directly in a 133 mathematical model calibrated using artificial mixtures. To achieve this goal, all the 134 135 previous studies used the Partial Least Squares Regression method (PLSR) (Table 1). However, more robust and non-parametric models, such as Support Vector Machine 136 (SVM), could usefully be tested be tested for this type of application, as they were 137 138 successfully used to derive soil properties, including clay and organic carbon content (Dotto 139 et al., 2017; Lucà et al., 2017; Stevens et al., 2013; Viscarra Rossel and Behrens, 2010). This 140 multivariate method seeks to identify an interpolation function using the kernel function, 141 adjusting the calibration data until simultaneously minimizing the size of the coefficients 142 and the prediction errors, in which data with non-linear patterns can be better represented 143 by the calibrated model (Ivanciuc, 2007).

In spectroscopic fingerprinting studies, there is also a knowledge gap regarding the use of 144 145 spectrum pre-processing techniques in spectroscopic modeling (Table 1), despite the fact 146 that several soil studies demonstrated that this step is extremely important when calibrating the models (Buddenbaum and Steffens, 2012; Dotto et al., 2017; Nawar et al., 147 2016). Several spectrum pre-processing techniques can be used for this purpose, such as 148 149 smoothing, Savitzky-Golay with 1st or 2nd derivative using a first or second order 150 polynomial, standard normal variate, multiplicative scatter correction and normalization. 151 However, so far, very few fingerprinting sediment tracing studies have reported whether

they used any type of spectral pre-processing. Most of them used only pre-processing
(Brosinsky et al., 2014b, 2014a; Chapkanski et al., 2019; Tiecher et al., 2017, 2016, 2015),
except for Ni et al. (2019), who used four pre-processing techniques, although they did not
compare them to each other.

In this context, the current research provides, to the best of our knowledge, the first 156 attempt to compare the outputs of different multivariate mathematical models (both 157 158 parametric and non-parametric) and preprocessing techniques of reflectance in UV-Vis, NIR 159 and MIR spectral bands used in combination or in isolation to predict the contribution of sediment sources at the catchment scale. Accordingly, the objectives of the study were (i) 160 161 to evaluate the accuracy of two multivariate methods for sediment source apportionment 162 (PLSR and SVM), (ii) to evaluate the effect of eight spectra preprocessing techniques, and 163 (iii) to evaluate the effect of using the information contained in the UV-Vis, NIR and MIR regions either separately or in combination. For this purpose, the estimated contribution 164 for each source using the spectroscopic models was also compared against the values 165 166 obtained by the conventional fingerprinting approach based on geochemical tracers

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#### 168 2. Materials and methods

#### 169 **2.1. Study site**

The Arvorezinha catchment is located in the northeastern part of the Rio Grande do Sul State, southern Brazil. Igneous rocks (basalts and rhyodacite) characterize the geology and the altitude varies from 580 to 730 meters. The upper third of the catchment has an undulating plateau relief with slopes up to 7%, and the middle and lower thirds of the catchment have a much steeper topography with slope gradients often exceeding 15%. The 175 climate is classified as Cfb (subtropical super-humid with no dry season and warm summer) according to Köppen (Alvares et al., 2013). The mean annual precipitation for the last 15 176 years (2002-2016) is 1938 mm with a mean erosivity of 9344 MJ mm ha<sup>-1</sup> h<sup>-1</sup> yr<sup>-1</sup> (Ramon, 177 2017). The main crop is tobacco grown in small farms. Corn, soybean, eucalyptus and 178 179 native forests are also found in the catchment. The landscape is characterized by short, 180 steep slopes with a strong hydrological connectivity between hillslopes and the drainage 181 network. The soil classes found in the catchment are Acrisols, Cambisols and Leptosols 182 (IUSS Working Group WRB, 2015). Inadequate soil management under agricultural land associated with limited water infiltration due absence of subsurface horizon or clayey B 183 horizon favor the formation of runoff that controls erosion dynamics in cropland. In 184 addition to erosion processes in cropland, the inadequate location and maintenance of 185 186 unpaved roads generate preferential pathways for runoff concentration accelerating erosion, and converting these roads into significant potential sources of sediment. The 187 188 catchment also shows signs of channel banks erosion due to the high flow energy observed during the events (flash flood) associated with the absence of riparian forest along several 189 190 river sections. Further details on the catchment can be found in Tiecher et al. (2015).

191 Soil samples were collected from the three main potential sources of sediment, including (i) 192 cropland surface, (ii) unpaved roads and (iii) stream channels. Cropland (n=20) and 193 unpaved road (n=10) samples were taken using a non-metallic trowel from the uppermost 194 layer (0–0.05 m). Stream channels (n=10) were sampled on exposed bank sites located 195 along the main river channel network. Each sample was composed of at least 10 subsamples collected in the vicinity of the sampling point (within a radius of approximately 196 197 10 m). Source sampling sites were selected based on visible signs of soil erosion and 198 hydrological connectivity as well as taking into account pedological variability. In order to

199 characterize the sediments transported in the drainage network, 29 suspended sediment samples were collected during nine rainfall events from October 2009 to July 2011 at the 200 201 catchment outlet, where the flow and sediment concentration are also continuously monitored, which allows the calculation of liquid and solid discharges, as well as the 202 203 catchment sediments yields. The sampled events cover the seasonal variability of land 204 cover in the drainage area, as well as the variations of river flow conditions. For high 205 magnitude events, several samples were collected to characterize the intra-event 206 variability (rise, peak and recession of the hydrograph). Sediment concentrations during monitored events ranged from 300 to 2000 mg L<sup>-1</sup>. To obtain a sufficient mass of 207 208 suspended sediment for conducting all analyses, samples were collected using a portable 209 continuous flow centrifuge (Alfie-500 Alfa Laval). All the source material and sediment 210 samples were oven-dried at 50 °C, gently disaggregated using a pestle and mortar, and 211 passed through a 63-µm mesh prior to laboratory analyses to investigate similar particle size-fractions for all the samples (suspended sediments and sediment sources). 212

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#### 214 **2.2. Artificial mixtures of sediment sources**

The samples of each potential sediment source were mixed in equal proportions in the laboratory to constitute a unique reference sample for the corresponding source. Then, those reference samples were mixed in 48 different weight proportions as presented in Fig. 2. These artificial mixtures containing different proportions of the three sources were then used to calibrate the multivariate mathematical models used to estimate the source contributions to the suspended sediment samples.

#### 222 **2.3. Spectral analyses**

UV-Vis diffuse reflectance spectra of samples were recorded at room temperature from 200 to 800 nm with a 1-nm step using a Cary 5000 UV–Vis–NIR spectrophotometer (Varian, Palo Alto, CA, USA). Samples were ground and loaded into a Harrick Praying Mantis diffuse reflectance accessory that uses elliptical mirrors. BaSO<sub>4</sub> was used as a 100% reflectance standard. Care was taken when adding the samples into the sample holder to avoid differences in sample packing and surface smoothness.

Near infrared (NIR) spectra were recorded in the range 10000–4000 cm<sup>-1</sup> (1000-2500 nm) using a Nicolet 26700 FTIR spectrometer (Waltham, Massachusetts, USA) in diffuse reflectance mode with an integrating sphere and a InGaAs detector with a resolution of 4 cm<sup>-1</sup> and 100 readings per spectrum.

233 Mid infrared (MIR) spectra were obtained in the range 400–4000 cm<sup>-1</sup> (2500-25000 nm) 234 using a Nicolet 510-FTIR (Thermo Electron Scientific, Madison, WI, USA) spectrometer in 235 diffuse reflectance mode with a resolution of 4 cm<sup>-1</sup> and 100 readings per spectrum. A 236 direct current of air was used (dry and without  $CO_2$ ) to eliminate  $CO_2$  and water from the 237 spectrometer in order not to interfere with scanning and obtaining the spectra.

UV-Vis, NIR and MIR spectral data were subjected to spectral pre-processing to remove physical variability due to light dispersion and to remove systematic variations of instrumental and environmental conditions in order to emphasize the characteristics of interest along the spectrum. Spectra without pre-processing constitute the "control treatment" (RAW). Eight spectral processing techniques commonly employed in chemometric studies (Rinnan et al., 2009) were tested to evaluate its effect on the calibration of spectroscopic models. 245 Pre-processing includes: (i) smoothing (SMO) from a convolution function using a 25 nm 246 mobile window, after initial testing to define the best search window; (ii) Savitzky-Golay (Savitzky and Golay, 1964) with 1st derivative using a first order polynomial (SGD1), with 25 247 nm search window, after initial testing to define the best search window; (iii) Savitzky-248 249 Golay (Savitzky and Golay, 1964) with 2nd derivative using a second order polynomial 250 (SGD2), with 25 nm search window. The 1st and 2nd derivates calculate the change of 251 reflectance in wavelength variation rate. This technique is widely used to remove baseline 252 shifts and highlight spectral features of interest; (iv) varied normal standard deviation (SNV) is used to remove spectral data dispersion caused by noise and different particle 253 sizes and consists of subtracting the mean and dividing it by the standard deviation 254 255 (spectrum - mean/standard deviation) of each spectrum individually; (v) multiplicative 256 scatter correction (MSC) is effective in minimizing baseline compensations and multiplicative effects; (vi) normalization (NOR) is the ratio of spectrum bands measured by 257 standard deviation (NOR); (vii) combination of NOR+SGD1; (viii) combination of 258 MSC+SGD1. These eight techniques can be divided into three groups according to the 259 260 objective and the mathematical approach employed. The first group includes only the 261 smoothing of the spectra, represented by the SMO. The second group is defined by the use of derivatives to remove baseline shifts and enhance spectral features, represented by 262 263 SGD1 and SGD2. The third group corresponds to techniques for spectral data normalization and dispersion corrections such as SNV, MSC and NOR. All pre-processing was performed 264 265 using the prospectr and clusterSim packages (Stevens and Ramirez-Lopez, 2020; Walesiak 266 and Dudek, 2020) R software (R Core Team, 2020).

#### 268 **2.4. Spectroscopic model development: calibration and validation**

The spectroscopic models were calibrated from the spectral signature of the 48 mixtures with different proportions of each sediment source (Figure 2). Spectral data were generated from the following regions of the electromagnetic spectrum: UV-Vis, NIR, MIR, UV-Vis+NIR, UV-Vis+MIR, NIR+MIR and UV-Vis+NIR-MIR. Two multivariate calibration methods were used to adjust the models, where the effect of the eight spectral preprocesses plus the raw spectrum (RAW) was tested. This totaled 126 spectroscopic models for each source, totaling 378 models (Figure 4).

276 Two multivariate methods with different approaches were selected to calibrate the 277 spectroscopic models: (i) Partial Least Squares Regression (PLSR) (R pls package (Mevik et al., 2016)) parametric technique widely used in spectroscopic modeling (Angelopoulou et 278 279 al., 2020; Dotto et al., 2018) and (ii) Support Vector Machines (SVM) (R e1071 package 280 (Meyer et al., 2019)), non-parametric technique. Methods with different approaches were 281 selected due to the occurrence of linear and non-linear correlations between the organomineral components of soil/sediment and the spectral variables (Viscarra Rossel and 282 283 Behrens, 2010). The PLSR model handles data sets containing many independent and highly 284 correlated variables, such as UV-Vis-NIR-MIR spectral data. PLSR analysis reduces large data 285 sets to a small number of uncorrelated orthogonal factors to minimize the sum of the squares of the predicted value errors (Varmuza and Filzmoser, 2009). The SVM model was 286 used with the kernel function, which separates the calibration data into hyperplanes and 287 seeks to establish correlations between the dependent and independent variables when 288 these have non-linear behavior (Ivanciuc, 2007). 289

290 The parameterization in the calibration of the PLSR method was: method = 'pls', resampling method = 'cross-validation 10 k-fold', and number of components = '.ncomp = seq(1, 20,291 292 1)'. For SVM the parameters were: method = 'svmLinear', resampling method = 'crossvalidation 10 k-fold', and Kernel parameters = 'Support Vector Machine with Linear Kernel'. 293 294 Each model was calibrated with 70% of the samples (n= 34) and validated with 30% of the 295 samples (n = 14). Both sets were randomly generated. To evaluate the accuracy of the models the following parameters were calculated: coefficient of determination (R<sup>2</sup>) 296 (Equation 1), bias (Equation 2) and mean square root of the prediction error (RMSE) 297 (Equation 3). 298

299

$$R^{2} = \frac{\sum_{i=1}^{n} (\hat{y}_{i} - \bar{y}_{i})^{2}}{\sum_{i=1}^{n} (y_{i} - \bar{y}_{i})^{2}}$$
(1)

301

300

302 bias = 
$$\frac{1}{n} \sum_{i=1}^{n} (\hat{y}_i - \bar{y}_i)$$
 (2)

303

RMSE = 
$$\sqrt{\frac{1}{n} \sum_{i=1}^{n} (\hat{y}_i - y_i)^2}$$
 (3)

304

305 where:  $\hat{y}$  = predicted value of each source;  $\bar{y}$  = observed mean value of each source in the 306 mixture; y = observed values of each source in the mixture; n = number of samples with i = 307 1, 2,..., n.

309 **2.5** Estimation of sediment source contributions by spectroscopic models and 310 independent validation comparing with the contribution results obtained with the 311 conventional geochemical approach

312 The contribution of sediment sources was estimated by spectroscopic models from an independent set of spectral reflectance data, which were obtained from 29 suspended 313 314 sediment samples. The predicted values were then compared with those obtained with the 315 geochemical tracers (Tiecher et al., 2015) and the bias<sub>sp</sub> error statistics (Equation 2) and 316 RMSE<sub>sp</sub> (Equation 3) of the proportion of sediment estimated for each source were calculated. This approach used the total concentration in various elements (Ag, As, Cr, Fe, 317 318 Mo, and P) estimated by ICP-OES after microwave assisted digestion with concentrated HCl and  $HNO_3$  in the 3:1 ratio (aqua regia). Detailed information regarding sediment sampling 319 320 and the statistical procedure used in the conventional geochemical approach can be found 321 in Tiecher et al. (2015).

Finally, data of quality of the models based on validation with artificial mixtures of sediment (RMSE<sub>v</sub>), and compared with the sediment contribution values obtained with geochemical tracers (RMSE<sub>sp</sub>) were entered in a conditional inference regression tree procedure to highlight the factors that most influenced the quality of the models.

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#### 327 **3. Results and discussion**

328 **3.1. Model calibration performance** 

The results of the validation of the 378 models are presented in Figure 5. In general, there is variation in the predictive behavior of the models depending on the type of spectral preprocessing, spectral range, multivariate method and sediment source considered (Figure 332 5). One the one hand, the calibrated models with SGD1, NOR+SGD1 and NOR preprocessing achieved a greater accuracy in predictions, differing significantly from that 333 obtained with other techniques (Figure 6a). Studies on organic carbon and soil clay content 334 335 predictions have also shown that it is possible to increase the accuracy of spectroscopic 336 models calibrated with SGD1 and NOR pre-processing (Dotto et al., 2018; Knox et al., 2015; 337 Moura-Bueno et al., 2019; Pinheiro et al., 2017; Vasques et al., 2008). On the other hand, 338 the models calibrated with preprocessed spectra using the SGD2, MSC, MSC+SGD1, SMO and SNV techniques presented the lowest performance among all spectral ranges and their 339 340 combinations in the three sediment sources (Figure 6a). These results agree with the 341 studies of Dotto et al. (2018) and Moura-Bueno et al. (2019) that showed a lower accuracy in organic carbon predictions of subtropical soils in southern Brazil when the SVM and PLSR 342 343 models are calibrated with Vis-NIR spectra submitted to the MSC and SNV techniques, 344 respectively.

The SDG2 technique stands out with a higher value of RMSE<sub>v</sub> observed in the predictions of 345 346 the three sources, differing from the other pre-processes (Figure 6a). Moreover, the SDG2 347 technique was the only pre-processing that resulted in a  $RMSE_v$  value higher than 15% and was also the only method with a RMSE<sub>v</sub> higher than that obtained for the spectra without 348 any type of pre-processing (RAW). Possibly, the SGD2 technique may be eliminating 349 350 features (predictor variables) of the spectra that are important for the prediction of 351 sediment sources. Nevertheless, several studies have used the 2nd derivative technique to 352 treat UV-Vis, NIR and MIR spectra (Tiecher et al., 2015, 2016, 2017) and MSC (Ni et al., 2019) and SNV (Chapkanski et al., 2019) techniques to treat MIR spectra, with the objective 353 to derive the characteristics of interest along the spectrum and to improve the predictions 354

of the models of sediment source contributions. However, as observed in the current
 research, these pre-processing methods do not appear to be the most promising.

357 There is still no consensus on how to define a priori the pre-processing techniques that will 358 produce predictive models with a greater accuracy. Each pre-processing method will behave differently (Rinnan et al., 2009) depending on the set of soil samples considered, or 359 360 as in the current research, on the intra-source spectral variation. Thus, ideally, it remains 361 necessary to perform preliminary tests with different pre-processing methods, especially in 362 river catchments with contrasted geologies and soil types, and when considering greater number of potential sediment sources and in larger basins, where the complexity is 363 364 greater. The content in organic matter, clay and iron oxides strongly influence the spectral behavior of soils (Moura-Bueno et al., 2019; Viscarra Rossel and Behrens, 2010) and 365 sediments (Figure 3) (Tiecher et al., 2017, 2016, 2015). Therefore, in a context of great 366 variability between sources, as when comparing sources with different geology and 367 mineralogy (Poulenard et al., 2009), or surface and subsurface sources with contrasting 368 369 carbon contents (Evrard et al., 2013), pre-processing techniques may result in limited 370 improvement of predictive models. However, they may have a greater potential to improve models in larger basins and in study areas characterized by sources showing homogeneous 371 carbon contents and mineralogical compositions. 372

The lowest variation and value in the RMSE<sub>v</sub> for the unpaved roads source in all spectral ranges and combinations (Figure 6c) compared to the other potential sediment sources are likely associated with the spectral behavior of this source. Among the three sources considered, unpaved roads are the most depleted in organic matter and the coarsest grainsized (Tiecher et al., 2019). Furthemore, it is composed of subsoil material that is richer in 378 2:1 clay mineral (Tiecher et al., 2016). In addition, qualitatively, organic matter from 379 unpaved roads contains a higher proportion of alkyl-benzene, and a lower proportion of polysaccharides and amino acids, compared to cropland and stream channels, which in 380 turn are sources that show organic matter of similar quality (Tiecher et al., 2015). The 381 lower organic matter content of the unpaved roads strongly limits masking effects on 382 383 spectral features related to iron oxides between 450-850 nm. The influence of organic 384 matter on the shape and albedo of the spectral curve along the entire UV-Vis+NIR 385 spectrum, with an emphasis on specific regions of Fe oxides, was reported in studies conducted in Brazil (Dalmolin et al., 2005; Galvao and Vitorello, 1998; Moura-Bueno et al., 386 2019) and worldwide (Ben-Dor, 1997). This results in lower spectral variation (Viscarra 387 Rossel and Behrens, 2010) compared with samples with higher organic matter content, 388 389 such as cropland and stream channels. This explains the higher RMSE<sub>v</sub> variation in the UV-Vis range among different pre-processing methods observed for these two sources 390 compared to the unpaved road source (Figure 5a). Consequently, the calibrated models for 391 392 unpaved roads achieve a greater accuracy in estimates. In addition, the unpaved road source is also richer in hematite oxides of Fe, with prominent features compared to crop 393 394 fields and stream channels (Tiecher et al., 2015). This shows that the composition of each source affects the spectral behavior of sediment differently in each spectral range (Figure 395 396 3) and, consequently, this affects in turn the predictive models calibrated for each source (Batistelli et al., 2018). These results are also in line with those observed in soil 397 398 spectroscopic modeling, in which variations in the organo-mineral composition of the soil 399 were shown to strongly influence the performance of predictive models (Araújo et al., 400 2014; Moura-Bueno et al., 2019; Wijewardane et al., 2016) with an emphasis on parametric 401 multivariate methods (Lucà et al., 2017; Ramirez-Lopez et al., 2013).

402 By comparing the spectral ranges, it is possible to note that the RMSE<sub>v</sub> has decreased in the 403 following order: UV-Vis > NIR > MIR (Fig. 6b). Good calibration results using MIR were also observed in previous sediment tracing studies (Collins et al., 2014; Evrard et al., 2013; Ni et 404 405 al., 2019; Poulenard et al., 2012). This may be related to an increased sensitivity and an 406 improved identification of functional groups of organic matter by MIR compared to UV-Vis 407 and NIR (Viscarra Rossel and Behrens, 2010). Studies show that spectroscopic models 408 calibrated with MIR region spectra have the potential to discriminate different fractions of organic matter (Knox et al., 2015) and sediment sources (Evrard et al., 2013; Ni et al., 409 2019). This also explains why, in the current research, the models calibrated with MIR 410 411 spectra and their combinations presented the lowest variation in the RMSE<sub>v</sub> of predictions among all pre-processing techniques (Figures 5c, 5e, 5f and g), in particular for the models 412 413 calibrated with the SVM method.

414 Our results show that there was no significant gain in terms of RMSE<sub>v</sub> when combining the UV-Vis+NIR, UV-Vis+MIR and NIR+MIR bands, but that RMSE<sub>v</sub> decreases significantly when 415 416 the three bands of the electromagnetic spectrum were combined in a single model (UV-Vis+NIR+MIR) (Figure 6b). The same result has been observed in studies performed to 417 estimate the concentration of elements in soil samples (Knox et al., 2015; Reeves, 2010; 418 419 Soriano-Disla et al., 2014). The increase observed in  $RMSE_v$  for models calibrated with UV-420 Vis range (Figure 6b) is a consequence of the lower UV-Vis spectral range (200-800 nm), 421 which does not allow distinguishing organic matter components (Viscarra Rossel and 422 Behrens, 2010), associated with the absence of NIR range, which is more sensitive to detect clay minerals. This implies a reduction in the predictive capacity of models calibrated with 423 UV-Vis only or with the absence of the NIR band. This behavior is observed for models 424 425 calibrated with UV-Vis+MIR range, where the absence of NIR range increased the 426 prediction error (RMSE<sub>v</sub>) of the estimates (Figure 6b). Although the RMSE<sub>v</sub> of the models combining Vis-UV+MIR are intermediate to the RMSE<sub>v</sub> of the models calibrated using these 427 spectral ranges separately (i.e. Vis-UV and MIR alone), it was expected that the errors 428 would be smaller when combining them. A possible explanation for this result may be the 429 430 overfitting caused by the high number of parameters and a relatively low number of 431 artificial mixture samples used for calibrating the models (48 in total). Therefore, the low 432 number of artificial mixture samples used to calibrate the models can be a limitation of this study, especially when combining different spectral ranges. Further efforts in future studies 433 should evaluate the effect of the number of artificial mixture samples on the overfitting of 434 435 spectroscopic models.

In summary, the explanation for this difference is associated with the interaction of UV-Vis, 436 437 NIR and MIR wavelength bands, which are bands basically related to color, particle size, type of minerals and organic matter, their chemical bonds and functional groups (Viscarra 438 Rossel and Behrens, 2010), resulting in gain of explanatory information of the spectral 439 440 variation of the data (Viscarra Rossel et al., 2006). Therefore, as sediment often consists of a heterogeneous mixture, there is a greater variation of these factors, resulting in a larger 441 amount of functional groups detected on their surface and, therefore, the combination of 442 several wavelengths, for example, UV-Vis+NIR+MIR, enhances the performance of 443 444 spectroscopic models to discriminate between contrasted sediment sources.

In general, most spectroscopic models had RMSE<sub>v</sub> values below 15% for the validation, and
in some cases, values below 5% were found (Figure 5). The models calibrated to MIR, UVVis+NIR, UV-Vis+MIR and NIR+MIR spectra achieved RMSE<sub>v</sub> values lower than 10% in the
three sediment sources (Figure 6b), and those calibrated for the UV-Vis+NIR+MIR spectral

range, presented values lower than 7% for crop fields and stream channels and equal to 5% for unpaved roads. These values are lower than the error of 15%, which is often considered to provide the acceptable level for sediment fingerprinting studies (Collins and Walling, 2002). This shows that the use of the spectroscopy technique employing the most appropriate spectral processing approaches, spectral bands and multivariate methods has a great potential for discriminating the contribution of sediment sources.

455 The lowest RMSE<sub>v</sub> values in the predictions for both multivariate methods (PLSR and SVM) were observed for unpaved roads, which differed significantly from the other potential 456 sources (Figure 6c). The models calibrated with the SVM method had a better performance 457 458 compared to the PLSR, especially for the crop fields source (Figure 6c). This is explained by 459 the greater ability of the SVM method to model non-linear relationships (Ivanciuc, 2007; 460 Viscarra Rossel and Behrens, 2010) compared to the PLSR method. In this case, the crop fields source presents a heterogeneous organo-mineral composition, mainly regarding the 461 quality of organic matter. In this scenario, non-linear relationships between the organo-462 463 mineral components and the spectral variables of the crop fields source predominate. In this case, SVM method is able to establish explanatory correlations of data variance, 464 resulting in more accurate estimates. This also explains why the lowest variations in 465 predictions are observed for the SVM method, which is shown to provide a more stable 466 467 method for the modeling of sediment sources, with an emphasis on sources with greater variations in organo-mineral composition. 468

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# 470 3.2. Comparing sediment source predictions by spectroscopy models with those obtained 471 with the conventional geochemical approach

472 The estimated sediment contribution of each source obtained from the best calibrated models for the UV-Vis, NIR, MIR spectral ranges and their combinations is shown in Figure 473 7. In general, the PLSR and SVM models calibrated with spectra of the three spectral bands 474 and their combinations submitted to SGD1, NOR, and NOR+SGD1 pre-processing have 475 476 achieved a greater accuracy in predicting sediment sources when compared to estimates 477 derived from the geochemical method tested by Tiecher et al. (2015) (Appendix 1). The 478 models calibrated with spectra processed by the SGD2 technique showed the lowest 479 predictive performance among all approaches (Appendix 1). These results corroborate those observed during the calibration and validation of the models (Figs. 5 and 6). 480

481 In all modeling approaches, estimates of PLSR and SVM models clearly indicate that the main contribution of sediment is supplied by the crop fields source, followed by stream 482 channels and unpaved roads (Figure 7), corroborating the results estimated by the 483 geochemical method (Tiecher et al., 2015). Good agreement is observed between the 484 contribution of the sources obtained with the geochemical tracer approach and that 485 486 estimated by the PLSR and SVM models when combining the UV-Vis+NIR+MIR spectral ranges, with the crop fields source contributions amounting to 57%, 62% and 55%, 487 respectively, followed by the unpaved roads sources, with 23%, 24% and 19%, and stream 488 channel 20%, 20% and 21%, respectively (Figure 7g). 489

In general, the highest RMSE<sub>sp</sub> values were observed for crop fields and stream channel source estimates and the lowest for unpaved roads (Appendix 1). The predictions of crop fields and stream channel sources by the models generated with the SVM method were better than with PLSR. For the unpaved roads source, the SVM and PLSR methods presented very close RMSE<sub>sp</sub> values. For example, for the SVM method, the model with the

most accurate estimates was SVM-SGD1-UV-Vis+NIR+MIR, with RMSEsp of 17.7, 16.4 and 495 13.9%, and bias<sub>sp</sub> of 2.0, -0.7 and -2.7% for crop fields, stream channel and unpaved roads, 496 respectively (Figure 8; Appendix 1). Accordingly, an overestimation of the crop fields 497 contribution to sediment was found, such as an underestimation of the stream channel and 498 499 thee unpaved roads source contributions. Despite the good accuracy of the models 500 calibrated and validated with the mixtures of proportions of each source of sediment 501 (Figure 8), there is still a considerable error compared to predictions made with 502 fingerprinting based on geochemical tracers. This can be partly attributed to the low number of sediment samples evaluated (n = 29). Future studies should use a larger number 503 504 of sediment samples in order to better understand this relationship. Nevertheless, the 505 magnitude of the contributions calculated by spectroscopy and geochemical approach is 506 very similar for each source (Figure 8).

507 For the PLSR method, the PLSR-NOR-UV-Vis+NIR+MIR model was more accurate, with higher RMSE<sub>sp</sub> for the crop fields (20.0%) and stream channel (18.7%) sources and also 508 509 larger under- (bias<sub>sp</sub> = -4.5%) and overestimations (bias<sub>sp</sub> = 3.6%), respectively, and the lower RMSE<sub>sp</sub> (14.5%) and bias<sub>sp</sub> (-0.1%) were observed for unpaved roads source. The 510 highest prediction errors observed for crop fields and stream channel sources are in 511 accordance with the findings of Tiecher et al. (2015), who used the PLSR method and NIR 512 513 spectral range to estimate sediment sources in the same study area. According to the 514 authors, the three sources is enriched in 1:1 kaolinite type minerals. However, an increase 515 in the abundance of 2:1 clay minerals may be observed in the following order: crop fields > stream channel >> unpaved roads, and an increase in the abundance of quartz may be 516 found in reverse order. In addition, crop fields and stream channels also have a higher 517 organic matter content than unpaved roads (Tiecher et al., 2016). Therefore, the closer 518

519 mineral composition and higher organic matter content of the crop fields and stream channels provide higher spectral variations, resulting in higher prediction errors (20.0 and 520 521 15.8% for crop fields and stream channel, respectively) than those obtained for the unpaved roads (15.5%) for the PLSR-NOR-UV-Vis+NIR+MIR model (Appendix 1). 522 523 Furthermore, this variation in the organo-mineral composition of the sources has a greater 524 effect on the prediction errors when looking at the estimates of the calibrated PLSR models 525 for the separate spectral ranges such as UV-Vis which showed RMSE<sub>sp</sub> of 23. 7, 26.6 and 526 15.7% for crop fields, stream channel and unpaved roads, respectively, for the PLSR-NOR-SGD1 model; and MIR with RMSE<sub>sp</sub> 25.8, 25.7 and 18.6% for crop fields, stream channel 527 528 and unpaved roads, respectively, for the PLSR-SGD1 model. This is because the UV-Vis and 529 MIR bands are less sensitive to clay mineral types (Viscarra Rossel and Behrens, 2010) than 530 the NIR spectral band, which identifies these constituents more clearly. Therefore, for the 531 PLSR-SGD1-NIR model, no major differences in RMSE<sub>sp</sub> values (23.8, 16.2 and 15.8% for crop fields, stream channel and unpaved roads, respectively) were observed with respect 532 to the PLSR-UV-Vis+NIR+MIR. 533

The difference between the multivariate methods is related to the statistical approach 534 followed in each method (PLSR - parametric and SVM - non-parametric). In this case, the 535 estimate of the contribution of the crop fields source obtained by the SVM-UV-536 537 Vis+NIR+MIR model presented a value very similar to that observed by the geochemical method (Figure 7g), indicating that this non-parametric model is more robust for the 538 539 spectroscopic modeling of this sediment source. Among the three sources considered, crop fields had greater spectral variation (Tiecher et al., 2015) and, therefore, concomitant 540 occurrence of linear and nonlinear correlations between spectral variables and sediment. 541 542 In this scenario, non-parametric methods present better adjustments in the models, especially SVM, which uses the kernel mathematical function to establish relationships between the dependent and independent variables, in which the model seeks to identify an interpolation function between the variables and creates support vectors (Ivanciuc, 2007). Studies have observed a similar behavior for the spectroscopic modeling of organic carbon content (Lucà et al., 2017; Viscarra Rossel and Behrens, 2010), and also that of exchangeable clay and calcium (Ramirez-Lopez et al., 2013) in soil samples with high spectral variations.

The results show that there is a difference in the performance of the calibrated 550 spectroscopic models with each spectral range and their combinations. For example, 551 552 models calibrated only with spectrum in the UV-Vis range have the highest error for both PLSR and SVM, with RMSE<sub>sp</sub> values > 22% for crop fields; RMSE<sub>sp</sub> > 26% for stream channels 553 and  $RMSE_{sp} > 20\%$  for unpaved roads (Appendix 1). By contrast, the lowest errors were 554 555 achieved in the UV-Vis+NIR+MIR ranges, where RMSE<sub>sp</sub> values were ~18% for crop fields and ~16% for stream channels and unpaved roads. This shows that when using narrower 556 557 spectrum bands there is a loss of information and consequently a loss of discriminating power of the models. However, the models that combined the three UV-Vis+NIR+MIR 558 spectral bands (Figure 7g) achieved a greater accuracy due to the better discrimination of 559 the inherent compositional characteristics of each source, as all major components that 560 561 may influence the spectral behavior (organic matter, clays and oxides) are taken into account in the spectra (Knox et al., 2015; Reeves, 2010; Viscarra Rossel et al., 2006). 562

563 Other differences in error metrics with respect to spectral ranges are observed by the 564 higher predictive capability of UV-Vis+NIR+MIR (Appendix 1) compared to models using 565 only UV-Vis+NIR range. This is in accordance with the findings of Reeves (2010) and Knox et 566 al. (2015) for soil organic carbon estimation. Furthermore, a study conducted by Bellon-Maurel and McBratney (2011) and Knox et al. (2015) showed that models developed to 567 quantify organic carbon content using MIR region data only produce slightly better results 568 than the UV-Vis+NIR region. However, this was not observed in the current research, 569 570 where models using only MIR showed a lower performance compared to that obtained 571 with UV-Vis+NIR (Appendix 1). It should be noted that in this study we are modeling sediment source contributions, which is very different from obtaining spectroscopic 572 573 estimates of elemental concentrations. Sediments consist of a mixture of particles with 574 different contents and types of clays, Fe oxides and organic matter. In this case, UV-Vis and 575 NIR spectral ranges have potential to discriminate between contrasted contents and types 576 of clays and Fe oxides (Viscarra Rossel and Behrens, 2010). They are therefore important 577 for discriminating between contrasted sediment sources, as already reported in the literature (Collins et al., 2014; Legout et al., 2013; Pulley and Rowntree, 2016; Tiecher et al., 578 2016). 579

580 Moreover, the estimates obtained by the models in the UV-Vis and MIR spectral ranges (Figure 7a, 7c, respectively) showed a greater dispersion between the PLSR and SVM 581 models and the three sediment sources. The same behavior is observed for the 582 combination UV-Vis+MIR (Figure 7e), particularly for crop fields. This may be attributed to 583 584 the interaction of UV-Vis and MIR wavelengths, which are bands related to the content and type of Fe oxides, and functional groups of organic matter, respectively (Viscarra Rossel 585 586 and Behrens, 2010). As sediment is a rather heterogeneous mixture, mainly supplied by crop fields, there is a greater compositional variation in these samples, which in this case, 587 the UV-Vis and MIR spectra are unable to capture. This observation interferes in the 588 589 correlations between sediment and spectral bands and, therefore, in the predictive power

590 of models. By contrast, the NIR spectral band resulted in estimates very close to those obtained by the geochemical method, with emphasis on the crop fields and unpaved roads 591 sources (Figure 7b). The explanation for this is due to the wavelengths corresponding to 592 the NIR region being able to jointly identify particle size, type of minerals and organic 593 matter, and supertons of chemical bonds and functional groups (Viscarra Rossel and 594 595 Behrens, 2010). Furthermore, the combinations of spectral bands with NIR (such as 596 NIR+MIR and UV-Vis+NIR+MIR) present the same tendency as that observed for NIR, i.e., 597 lower amplitude in estimates (Figure 7f, 7g). This shows that the NIR spectral range is the most important region of the electromagnetic spectrum for building spectroscopic models 598 for estimating sediment source contributions. This confirms previous findings obtained in 599 600 studies conducted at different locations around the world, which showed the good 601 performance of models that use data derived from the NIR spectral range to estimate sediment source contributions (Collins et al., 2014; Tiecher et al., 2016). 602

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#### 604 **3.3. Assessing the quality of the models**

605 It is important to note that in all models tested here, the contribution of each source is 606 estimated independently, i.e. each model estimates the proportion of a source 607 independently of the other two sources. Therefore, the sum of the estimates generated for 608 each source can provide a good indicator of model quality. In this case, it is understood 609 that models with the sum of contributions from sediment sources closer to 100% are 610 better (Legout et al., 2013). Figure 8 shows this comparison, where it is observed that 611 under and overestimation (ranging from 90 to 132% - Figure 9a) of the sum of the sources 612 for some approaches occur. Regarding the spectral pre-processing techniques,

613 overestimation occurs for the models calibrated for both multivariate methods (PLSR and SVM) with raw spectra (RAW) and submitted to SMO, SGD2, SNV, MSC and NOR techniques 614 (Figure 9a). Moreover, a slight underestimation is observed for MSC+SGD1. It is noted that 615 616 RAW and SMO spectra presented the highest overestimates and showed significant 617 differences compared to the others, indicating that the absence of preprocessing of 618 spectral data and/or only the smoothing provided less accurate estimates of source 619 proportions. It is also possible to note that the NOR+SGD1 and SGD1 techniques presented 620 the lowest variations in the sum of the sediment source contributions (Figure 8a). This 621 corroborates the best performance observed for these models during the validation step 622 (Figure 6a). The models that reached values closer to 100% were PLSR-SGD1 (100.6%) and SVM-SGD1 (100.1%) (Figure 9a). 623

624 Regarding the spectral ranges, it is noted that the models calibrated with the combinations 625 NIR+MIR and UV-Vis+NIR+MIR presented the values closest to 100%, differing significantly from the others (Figure 8b). The model calibrated with the SVM method and the 626 627 combination of UV-Vis+NIR+MIR presented the values closest to 100% (100.1%) (Figure 8b). Therefore, the results observed in Figure 8 corroborate those discussed in the section 628 dealing with the accuracy of the model validation and in the estimates of the sediment 629 630 source contributions. In addition, future studies should address the use of spectral variable 631 selection algorithms. This strategy has shown the potential to improve spectroscopic estimates (Xiaobo et al., 2010; Gomes et al., 2013; Hong et al., 2020). Additionally, research 632 633 employing the use of two-dimensional correlation (Hong et al., 2018) to identify regions or bands most correlated with different sediment sources can be a promising approach. 634

635 Finally, the decision tree analysis shows that the quality of calibration of spectroscopic models depends primarily on the spectral preprocessing technique, and secondarily on the 636 spectral band, and that the sediment source has little or no influence (Figure 10b). It is 637 evident that pre-processing with SGD2 and UV-Vis spectral band always result in higher 638 639 RMSE<sub>v</sub> values. However, when comparing the quality of the models based on the estimates 640 of sediment sources obtained with geochemical tracers (RMSE<sub>sp</sub>, Figure 10b), the spectral 641 band and sediment source is of greater importance, and NIR range or its combination with 642 the other spectral ranges result in contributions that are more similar to those obtained 643 with the geochemical approach.

#### 644 **4. Conclusions**

The current research demonstrated the great potential to improve the estimation of the 645 646 sediment source contributions using spectroscopy when using adequate spectral pre-647 processing technique, multivariate method, and spectral range. In general, the non-648 parametric support vector machine (SVM) model was more robust than the partial last square regression (PLSR), especially to estimate the contribution of sediment sources with 649 high organo-mineral variations, such as the crop fields source. For both models tested 650 651 (PLSR and SVM), a better performance was obtained using Savitzky-Golay spectral pre-652 processing techniques with 1<sup>st</sup> derivative (SGD1), normalization (NOR) and combining NOR+SGD1. Furthermore, it was verified that the combination of the three spectral ranges 653 of the electromagnetic spectrum tested (UV-Vis, NIR and MIR) enhanced the performance 654 655 of the spectroscopic models, resulting in lower errors in the predictions of the sediment source contributions. This is due to the sum of different information contained in each 656 657 spectral range related to the organic and mineral composition of each sediment source.

Despite the good accuracy of the models calibrated and validated with the mixtures, significant errors remain when comparing sediment source contributions c to the results obtained with the conventional sediment fingerprinting method based on geochemical tracers. Nevertheless, the magnitude of the contributions calculated by spectroscopy and geochemical approaches remains very similar for all sources. Efforts should be done in future studies to validate these findings in larger catchments as well as in sites where more potential sediment sources may supply material to the river systems.

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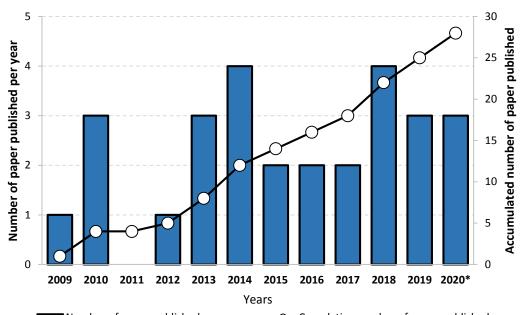
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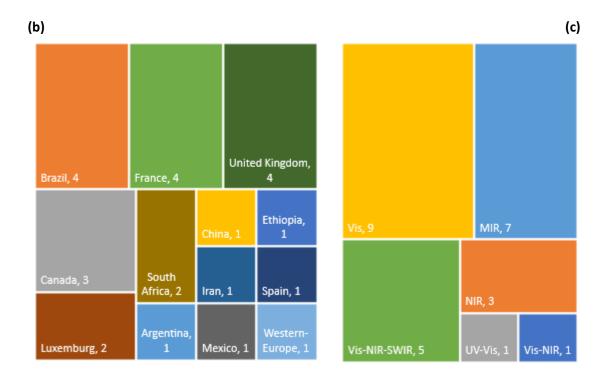
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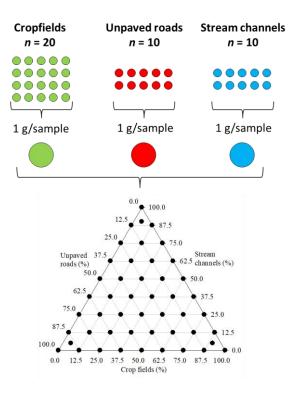
## 2 Figure 1.

Number of published scientific articles per year and cumulative number of published articles
 using spectroscopy to trace sediment sources for the period 2009-2020 (a), relative distribution

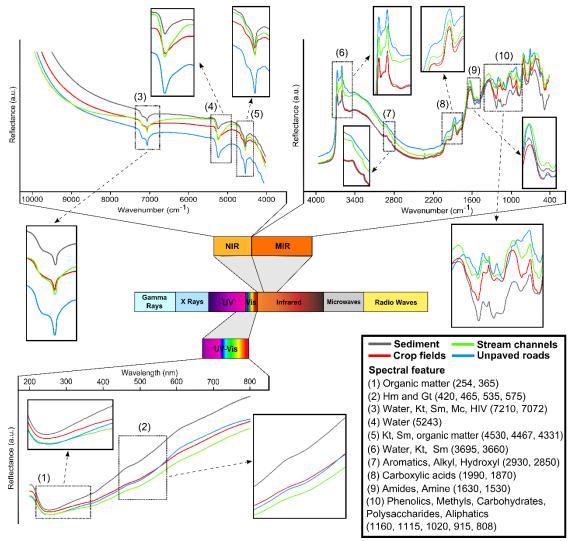
5 of spectroscopic fingerprinting studies by country (b) and spectral range (c). \*Until 15<sup>th</sup> June

6 2020. UV, ultraviolet. Vis, visible. NIR, near infrared. SWIR, short-wave infrared. MIR, mid

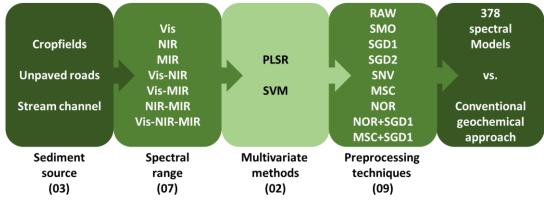
7 infrared.



- 8 9 **Figure 2.**
- 10 Ternary diagram with the position of the experimental mixtures prepared for the calibration
- 11 and validation of the spectroscopic-models.
- 12

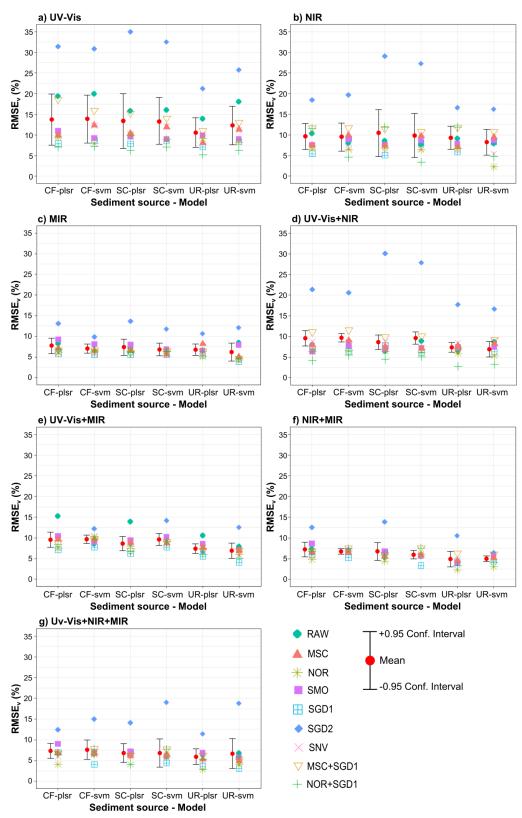


- 13
- 14 **Figure 3.**
- 15 Characterization of the main spectral features found in the UV-Vis, NIR and MIR ranges for the
- 16 suspended sediment and potential sediment sources in the Arvorezinha catchment.



**Figure 4.** 

19 Schematic construction of predictive models of the spectral signature of sediment sources.

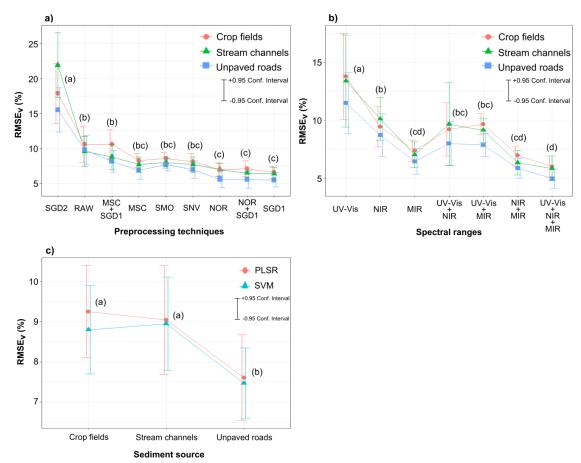




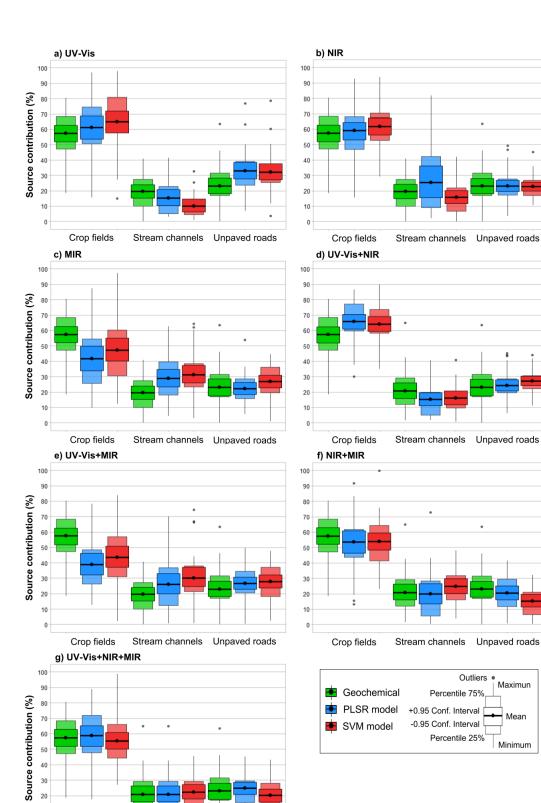
21 Figure 5.

Performance in validating Partial Last Square Regression (PLSR) and Support Vector Machine (SVM) prediction models from raw spectral data and combined with the eight spectral preprocessing techniques for the sediment sources including stream channels (SC), unpaved roads (UR) and surface of crop fields (CF). RAW - raw spectral; SMO - smoothing; SGD1 - Savitzky-Golay with 1<sup>st</sup> derivative; SGD2 - Savitzky-Golay with 2<sup>nd</sup> derivative; SNV - varied standard deviation

27 correction; MSC - multiplicative scatter correction; NOR - normalization by standard deviation.



- 28
- 29 Figure 6.
- 30 Mean values of the RMSE<sub>v</sub> error statistic of the validation of the prediction models of the three
- 31 sediment sources in relation to (a) eight spectral pre-processing techniques; (b) spectral bands
- 32 and their combinations; (c) Partial Last Square Regression PLSR and Support Vector Machine -
- 33 SVM. Means followed by the same letter do not differ by the Tukey's test at p<0.05.





35 **Figure 7.** 

10 0

Crop fields

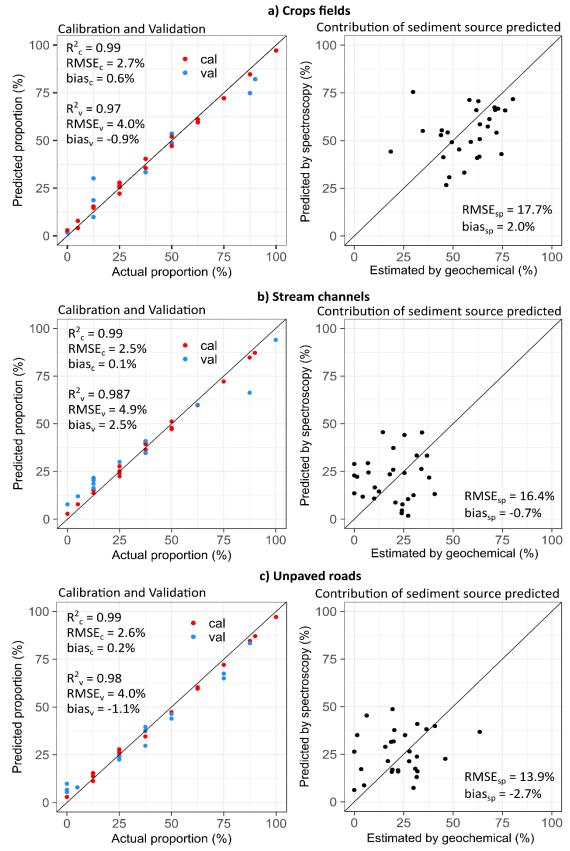
36 Boxplot of the contribution of sediment sources estimated by the different approaches for the

Stream channels Unpaved roads

37 29 sediment samples. The estimates are derived from the models that presented the highest

38 accuracy among all processing and spectral range combinations for the PLSR and SVM methods

- 39 (UV-Vis = SGD1, NIR = NOR+SGD1, MIR = SGD1, UV-Vis+NIR = NOR+SGD1, UV-Vis+MIR = SGD1,
- 40 NIR+MIR = NOR+SGD1, UV-Vis+NIR+MIR = SGD1).

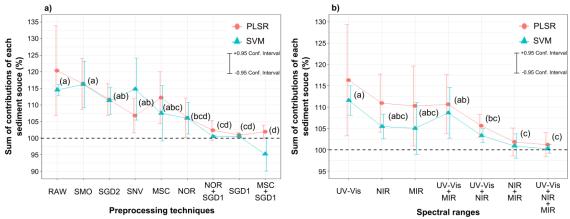


## 41 42 Figure 8.

43 Scatter plot of the validation and calibration of the best model (SVM-SGD1-UV-Vis + NIR + MIR)

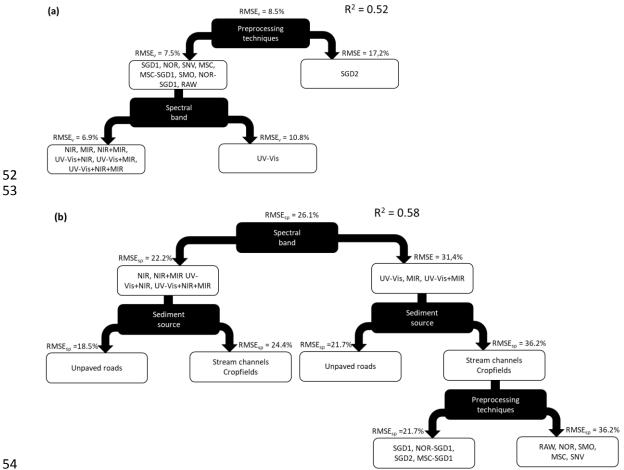
44 for mixtures the three sediment sources and the predicted contribution values by the spectra

45 and observed by the geochemical method.



## 46 47 **Figure 9.**

- 48 Sum of sediment source contributions comparing Partial Last Square Regression PLSR and
- 49 Support Vector Machine -SVM multivariate methods for each pre-processing technique (a) and
- 50 for each spectral range and their respective combinations (b). Means followed by the same letter
- 51 do not differ according to the Tukey's test at p<0.05. The dotted line represents 100%.



## 55 **Figure 10.**

- 56 Conditional inference tree analysis evaluating the factors that most affect the quality of the
- 57 models based on validation with artificial mixtures of sediment (RMSEv a), and compared with
- 58 the sediment contribution values obtained with geochemical tracers (RMSEsp -b).