Plutonium in the environment: key factors related to impact assessment in case of an accidental atmospheric release


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Plutonium in the environment: key factors related to impact assessment in case of an accidental atmospheric release


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Summary. This paper deals with plutonium and key factors related to impact assessment. It is based on recent work performed by CEA which summarize the main features of plutonium behaviour from sources inside installations to the environment and man, and to report current knowledge on the different parameters used in models for environmental and radiological impact assessment. These key factors are illustrated through a case study based on an accidental atmospheric release of Pu in a nuclear facility.

1. Introduction

Plutonium is a raw material of the nuclear industry and exists in a large variety of compounds.

The objective of present work is to illustrate through a case study, the important factors influencing the results of a radiological impact assessment study. This work is based on a recent report performed by CEA [1] which intends to present, discuss and summarize the main features of plutonium behaviour from sources inside installations to the environment and man. Transfer and behaviour in man and effects on health are also presented. The objective of this report is to gather scientific information useful for decision makers in case of accident or for regulation purposes.

The case study selected here is based on an accidental atmospheric release of Pu in a nuclear facility. The key factors which influence the radiological assessment study is discussed in the following sections.

2. Case study

A fire in a glove box containing 1 kg of heavy metal is considered. A basic safety calculation would consider the emission of 1% of 239Pu, and according to the default option selected by ICRP for populations [2], in the form of aerosol particles of 1 µm diameter with moderate solubility (M).

On this basis, keeping in mind the simplification objective as expressed just before, the exposure at 1 km from the emission point, for a release of 1 GBq of 239Pu will be of 1 mSv all pathways taken into account [1]. The exposure pathway is for 95% inhalation, for 4% ingestion of leaf vegetables and fruits, and for 1% ingestion of cereals. Exposure to children is less than that to adults and does not imply supplementary protection. For the case under interest, the release of 10 g of 239Pu would be of $2.3 \times 10^{10}$ Bq (or 20 GBq if oxygen mass is taken into account). The exposure at 48 h would be then above the intervention level of 10 mSv and would imply a confinement of populations. Protective actions relative to food consumption and trade restriction would also be useful and would concern about the same area.

Nature, isotopic composition and physico-chemical form

First, it is important to check that the accident deals with a PuO2 powder and not a MOX powder. Indeed, the 239Pu activity of ten grams of powder would be relatively smaller in the case of a fresh MOX powder containing 11.6% plutonium in mass (Table 1), that is $1.4 \times 10^8$ Bq of 239Pu per gram of powder.

Second, if 239Pu is the major isotope in terms of mass, it is not the case in terms of activity. In the case of MOX (Table 1), 238Pu activity is about twenty times higher than the activity of 239Pu with an equivalent inhalation dose coefficient. The 241Pu activity ($\beta$ emitter) is also higher ($5 \times 10^{10}$ Bq g$^{-1}$ of powder) but with an inhalation dose coefficient 50 times lower. An important point is the case of 241Am produced by filiation of 241Pu which must be taken into account.

Hence if we consider Pu isotopy of MOX 11.6%, the sum of activities of alpha emitters (which have relatively close dose coefficients) and also 2% of 241Pu activity, the toxicity by inhalation of one gram of this plutonium is 15 times higher than that of one gram of pure 239Pu. On this point
Table 1. Isotopic composition of fresh MOX 11.6% in mass, activity and dose.

<table>
<thead>
<tr>
<th>Isotopic composition</th>
<th>Specific activity Bq/g</th>
<th>Isotopic composition in activity</th>
<th>Activity of 1 g of Pu</th>
<th>239Pu equivalence factor in dose</th>
<th>Relative contribution in dose</th>
</tr>
</thead>
<tbody>
<tr>
<td>Pu238</td>
<td>3.6%</td>
<td>6.3 × 10^11</td>
<td>5.1%</td>
<td>2.3 × 10^10</td>
<td>1</td>
</tr>
<tr>
<td>Pu239</td>
<td>50.8%</td>
<td>2.3 × 10^10</td>
<td>0.26%</td>
<td>1.2 × 10^9</td>
<td>1</td>
</tr>
<tr>
<td>Pu240</td>
<td>24.9%</td>
<td>8.3 × 10^9</td>
<td>0.46%</td>
<td>2.1 × 10^7</td>
<td>1</td>
</tr>
<tr>
<td>Pu241</td>
<td>11.2%</td>
<td>3.8 × 10^8</td>
<td>94%</td>
<td>4.2 × 10^5</td>
<td>0.02</td>
</tr>
<tr>
<td>Am241</td>
<td>1.1%</td>
<td>1.3 × 10^5</td>
<td>0.33%</td>
<td>1.5 × 10^3</td>
<td>0.8</td>
</tr>
<tr>
<td>Pu242</td>
<td>8.4%</td>
<td>1.5 × 10^4</td>
<td>0.0%</td>
<td>1.3 × 10^2</td>
<td>1</td>
</tr>
<tr>
<td>Total</td>
<td>100%</td>
<td>100%</td>
<td>4.5 × 10^3</td>
<td></td>
<td>31–100%</td>
</tr>
</tbody>
</table>

there is a source of underestimation of the impact which could be important. Generally, the exposure to plutonium is mainly due to 238 and 241 isotopes and then to 239 isotope. The contribution of 236Pu is more variable as a function of plutonium age.

The third point concerns the physico-chemical form and the corresponding solubility (moderate M or slow S). The ratio between the S form for the dioxide and the M form taken by default varies between 3 and 5 according to the isotopes. An overestimation could then be made which could be avoided.

Granulometry

The last point concerns the size of released particles. This aspect is particularly important since it appears at different levels in the calculation. This parameter could also be related to the physico-chemical form.

Firstly, the particle size is involved in the resuspension process. In our example, a fire with destruction of the glove box could lead to a resuspension of 5 × 10^-3 of the powder activity (relatively close of the 1% used just above). But the inhalable fraction would be only about 10% of the resuspension and would induce an exposure by inhalation at least 10 times lower. It is interesting to note that this phenomena is even more important for nitrate or fluoride forms for which the inhalable fraction would be only about 10% of the resuspension. But the second point is the inhalable aerosol part varies from 15% at origin to 60% at 30 km of distance [1].

Fig. 2 shows the variation of the mean dose coefficient with the distance taking into account the granulometry evolution for different meteorological conditions, characterized by the stability conditions (St = stable, Un = unstable) and the wind velocity. This factor increases with the distance downwind since small particles become predominant. The curves shapes can be explained as follow. First, there is a conjugation of the S physico-chemical form selected and the important fraction (in activity) of size particle class (10 µm) in the content of inhalable particles. This leads to a mean dose factor five times lower. The second point is the importance at short distance of non inhalable particles. The separation of particles is performed more or less quickly according to the meteorological conditions. It should be noticed that the dose factor at short distance is 25 times less than the standard dose per unit intake coefficient (DPUI).
Contamination of vegetal and animal productions

The previous analysis show that the granulometry has a multiple influence on the main exposure, namely the inhalation. The exposure is then principally dependant of the activity fraction released as inhalable form. According to the inhalation pathway, particles with a diameter higher than 10 µm could be neglected. But the importance of captation of bigger particles (with a diameter higher than 10 or 100 µm) for the food pathway has to be studied. Indeed, the deposition velocity on ground (5 × 10⁻³ m/s) for inhalable particles is multiplied by a factor 20 for particles of 70 µm, and by a factor of 200 for particles of 200–300 µm (aerodynamic diameter). Ingestion then becomes the main pathway even with a dose factor lower that the inhalation dose factor.

Concerning the air-vegetable transfer factors and the biological decrease of deposit on vegetables [4], it appears that captation is lower for particles of some hundred µm than for inhalable particles by a factor of around 3. Captation is negligible for higher size particles. Granulometry is not a key factor for deposit captation by rain water. The losses of foliar deposit of aerosols are relatively rapid during the first days without insofar as being important. They are limited to a factor 2 or 3 of reduction after 10 to 15 d and then remain stable. Afterward, only the plant growth still reduces the concentration of edible parts.

The calculation described in [1] with a granulometry of 1 µm (and a sedimentation rate of 5 × 10⁻³ m/s) shows that the food trade limit could be overtaken up to a distance of 1 km for a release of around 1 g of ²³⁹Pu or also around 3 GBq of ²³⁹Pu. In our example the equivalent mass activity is 15 times higher. Moreover considering that (i) the Pu mass released is 5 g, (ii) the mean deposit rate at 1 km is 5 × 10⁻⁵ m/s and (iii) the vegetable captation is 3 times less, the trade limit would be exceeded by a factor of 200 which then increases the intervention area up to a distance of around 10 km.

The other exposure pathways are minor because plutonium is a non-metabolized element. There is no significant transfer from soil, no transfer between leaves and fruits or roots, and transfers to animal products (meat and milk) are relatively low. The resuspension of soil particle has been in the past considered to be a significant pathway. Analysis of Chernobyl accident, however, has shown that this is of minor importance.

The above analyses of the scenario lead to a better understanding of what could be the reality and the needs of intervention in case of an accident:

- First of all, the toxicity of one gram of plutonium largely depends on its isotopy. In the retained example, it is 15 times higher to one gram of pure ²³⁹Pu.
- Secondly, the granulometry of released particles plays a major role in the relative importance of the different exposure ways. For a release of fine particles (reference case since there is a filtration level), inhalation is the dominant exposure pathway. However, in the absence of filtration, the release could be higher with particles of greater sizes. The ingestion pathway may then become the most important. Under the conditions in our example, the inhalation exposure would remain less than 10 mSv in the early phase of the accident, out of the center, and would not imply confinement measures. But the food restriction intervention in the later phase, for leaf vegetables, would concerns an area of 10 km length.
- Thirdly, in both cases (with or without filtration), the consequences of an accidental Pu release are limited in time, the pathways playing a role at long term being secondary.
- Finally, results are insensitive to age and intervention measures should be applied (or not) to the whole population. The fetus transfer or through maternal milk transfer are also relatively low.

Relative aspects to exposure and risks

These points are rarely discussed in impact studies. The dose coefficients (dose per unit intake) are calculated in a standard way with a density of 3 kg L⁻¹ and a standard deviation of 2.5 kg L⁻¹. On these bases, a PuO₂ aerosol contains 420 particles per Bq [5]. In reality, the density is near to 9 kg L⁻¹ and the standard deviation to 1.7 kg L⁻¹. Under these conditions the particle number would be much lower, around 30 particles per Bq. Uncertainties on these dose factors increase linearly with the diminution of particle number.

It is interesting to notice that uncertainties on effective doses delivered to the population are low. On the contrary, they are much higher at the individual scale and insofar as the particle number is small.

Concerning the uncertainties on the risk, the spatial dose distribution could be very heterogeneous. The studies show that the presence of hot spots is systematically associated to the appearance of fibroses sources, which decreases the delivered doses to cell targets. Hence in the case of insoluble compounds of high specific activity, calculated equivalent doses according to the ICRP recommendations would be highly overestimated if they are assumed to reflect the cancer risk.

3. Conclusions

This study performed on the basis of a complete report [1] confirms information brought by [6] for respirable aerosols for fresh MOX powders. Nevertheless it shows that, at least
in some cases, when measured parameters are used, the impact of an accident could be very different from what is described by standard theoretical assessments. Details on the nature of plutonium handled in the facility, its isotopy and granulometry (in case of powders) must be known by the nuclear operators to accurately assess the impact. It is worth noting that because of the safety arrangements related to criticality risk, the inventories that could be concerned in a laboratory are necessarily limited.

References


